

RESEARCH ARTICLE

Evenness mediates the global relationship between forest productivity and richness

Iris Hordijk¹  | Daniel S. Maynard¹ | Simon P. Hart² | Mo Lidong¹ | Hans ter Steege^{3,4} | Jingjing Liang⁵ | Sergio de-Miguel^{6,7} | Gert-Jan Nabuurs⁸ | Peter B. Reich^{9,10} | Meinrad Abegg¹¹ | C. Yves Adou Yao¹² | Giorgio Alberti^{13,14} | Angelica M. Almeyda Zambrano¹⁵ | Braulio V. Alvarado¹⁶ | Alvarez-Davila Esteban¹⁷ | Patricia Alvarez-Loayza¹⁸ | Luciana F. Alves¹⁹ | Christian Ammer²⁰ | Clara Antón-Fernández²¹ | Alejandro Araujo-Murakami²² | Luzmila Arroyo²² | Valerio Avitabile²³ | Gerardo A. Aymard C²⁴ | Timothy Baker²⁵ | Radomir Bałazy²⁶ | Olaf Banki⁴ | Jorcely Barroso²⁷ | Meredith L. Bastian^{28,29} | Jean-Francois Bastin³⁰  | Luca Birigazzi³¹ | Philippe Birnbaum³² | Robert Bitariho³³ | Pascal Boeckx³⁴ | Frans Bongers⁸ | Olivier Bouriaud³⁵ | Pedro H. S. Brancalion³⁶ | Susanne Brandl³⁷ | Roel Brienen²⁵ | Eben N. Broadbent³⁸ | Helge Bruelheide^{39,40}  | Filippo Bussotti⁴¹ | Roberto Cazzolla Gatti⁴²  | Ricardo G. César³⁶ | Goran Cesljar⁴³ | Robin Chazdon^{44,45} | Han Y. H. Chen⁴⁶  | Chelsea Chisholm¹ | Emil Cienciala^{47,48}  | Connie J. Clark⁴⁹ | David B. Clark⁵⁰ | Gabriel Colletta⁵¹ | David Coomes⁵²  | Fernando Cornejo Valverde⁵³ | Jose J. Corral-Rivas⁵⁴ | Philip Crim^{55,56} | Jonathan Cumming⁵⁶ | Selvadurai Dayanandan⁵⁷ | André L. de Gasper⁵⁸ | Mathieu Decuyper^{8,59}  | Géraldine Derroire⁶⁰ | Ben DeVries⁶¹ | Ilija Djordjevic⁶² | Amaral Iêda⁶³ | Aurélie Dourdain⁶⁰ | Engone Obiang Nestor Laurier⁶⁴ | Brian Enquist^{65,66}  | Teresa Eyre⁶⁶ | Adandé Belarmain Fandohan⁶⁷ | Tom M. Fayle⁶⁸  | Leandro V. Ferreira⁶⁹ | Ted R. Feldpausch⁷⁰ | Leena Finér⁷¹ | Markus Fischer⁷² | Christine Fletcher⁷³ | Lorenzo Frizzera⁷⁴ | Javier G. P. Gamarra⁷⁵ | Damiano Gianelle⁷⁴ | Henry B. Glick⁷⁶  | David Harris⁷⁷ | Andrew Hector⁷⁸  | Andreas Hemp⁷⁹ | Geerten Hengeveld⁸ | Bruno Héroult^{80,81}  | John Herbohn⁴⁵ | Annika Hillers^{82,83} | Eurídice N. Honorio Coronado⁸⁴ | Cang Hui^{85,86} | Hyunkook Cho⁸⁷ | Thomas Ibanez⁸⁸ | Il Bin Jung⁸⁷ | Nobuo Imai⁸⁹ | Andrzej M. Jagodzinski^{90,91}  | Bogdan Jaroszewicz⁹² | Vivian Johanssen⁹³ | Carlos A. Joly⁹⁴ | Tommaso Jucker⁹⁵ | Viktor Karminov⁹⁶ | Kuswata Kartawinata¹⁸ | Elizabeth Kearsley⁹⁷  | David Kenfack⁹⁸ | Deborah Kennard⁹⁹ | Sebastian Kepfer-Rojas⁹³  | Gunnar Keppel¹⁰⁰  | Mohammed Latif Khan¹⁰¹ | Timothy Killeen²² | Hyun Seok Kim^{102,103,104,105} | Kanehiro Kitayama¹⁰⁶ | Michael Köhl¹⁰⁷ | Henn Korjus¹⁰⁸  | Florian Kraxner¹⁰⁹ | Diana Laarmann¹¹⁰ |

This is an open access article under the terms of the [Creative Commons Attribution-NonCommercial-NoDerivs](https://creativecommons.org/licenses/by-nc-nd/4.0/) License, which permits use and distribution in any medium, provided the original work is properly cited, the use is non-commercial and no modifications or adaptations are made.

© 2023 The Authors. *Journal of Ecology* published by John Wiley & Sons Ltd on behalf of British Ecological Society.

Mait Lang¹⁰⁸  | Simon Lewis^{25,110} | Huicui Lu¹¹¹  | Natalia Lukina¹¹² |
 Brian Maitner⁶⁵  | Yadvinder Malhi¹¹³ | Eric Marcon¹¹⁴ | Beatriz Schwantes Marimon¹¹⁵ |
 Ben Hur Marimon-Junior¹¹⁵ | Andrew Robert Marshall^{45,116,117} | Emanuel Martin¹¹⁸ |
 Olga Martynenko⁹⁶ | Jorge A. Meave¹¹⁹  | Omar Melo-Cruz¹²⁰ | Casimiro Mendoza¹²¹ |
 Cory Merow⁴⁴ | Miscicki Stanislaw¹²² | Abel Monteagudo Mendoza^{123,124} |
 Vanessa Moreno³⁶ | Sharif A. Mukul^{45,125}  | Philip Mundhenk¹⁰⁷ | Maria G. Nava-Miranda¹²⁶ |
 David Neill¹²⁷ | Victor Neldner¹²⁸ | Radovan Nevenic⁶² | Michael Ngugi⁶⁶ |
 Pascal A. Niklaus¹²⁹ | Jacek Oleksyn⁹⁰ | Petr Ontikov⁹⁶ | Edgar Ortiz-Malavasi¹⁶ |
 Yude Pan¹³⁰ | Alain Paquette¹³¹ | Alexander Parada-Gutierrez²² | Elena Parfenova¹³² |
 Minjee Park^{5,102} | Marc Parren¹³³ | Narayanaswamy Parthasarathy¹³⁴  | Pablo L. Peri¹³⁵ |
 Sebastian Pfautsch¹³⁶ | Oliver L. Phillips²⁵ | Nicolas Picard¹³⁷ | Maria Teresa Piedade¹³⁸ |
 Daniel Piotto¹³⁹ | Nigel C. A. Pitman¹⁸ | Irina Polo¹⁴⁰ | Lourens Poorter⁸ | Axel
 Dalberg Poulsen⁷⁷ | John R. Poulsen⁴⁹  | Hans Pretzsch¹⁴¹ | Freddy Ramirez Arevalo¹⁴² |
 Zorayda Restrepo-Correa¹⁴³ | Mirco Rodeghiero^{74,144} | Samir Rolim¹³⁹ |
 Anand Roopsind¹⁴⁵ | Francesco Rovero^{146,147} | Ervan Rutishauser¹⁴⁸ | Purabi Saikia¹⁴⁹  |
 Christian Salas-Eljatib^{150,151,152}  | Peter Schall²⁰  | Dmitry Schepaschenko¹⁰⁹ |
 Michael Scherer-Lorenzen¹⁵³ | Bernhard Schmid¹⁵⁴  | Jochen Schöngart¹³⁸ |
 Eric B. Searle¹³¹ | Vladimír Šeβeň¹⁵⁵ | Josep M. Serra-Diaz^{156,157} | Douglas Sheil^{8,158}  |
 Anatoly Shvidenko¹⁰⁹ | Javier Silva-Espejo¹⁵⁹ | Marcos Silveira¹⁶⁰ | James Singh¹⁶¹ |
 Plinio Sist⁸⁰ | Ferry Slik¹⁶² | Bonaventure Sonké¹⁶³ | Alexandre F. Souza¹⁶⁴ |
 Krzysztof Stereńczak²⁵  | Jens-Christian Svenning^{156,165} | Miroslav Svoboda¹⁶⁶ |
 Ben Swanepoel¹⁶⁷ | Natalia Targhetta¹³⁷ | Nadja Tchebakova¹³¹ | Raquel Thomas¹⁶⁸ |
 Elena Tikhonova¹¹¹ | Peter Umunay⁷⁵ | Vladimir Usoltsev¹⁶⁹ | Renato Valencia¹⁷⁰ |
 Fernando Valladares¹⁷¹ | Fons van der Plas¹⁷² | Do Van Tran¹⁷³ | Michael E. Van Nuland¹⁷⁴  |
 Rodolfo Vasquez Martinez¹²² | Hans Verbeek⁹⁶ | Helder Viana^{175,176} |
 Alexander C. Vibrans^{57,177} | Simone Vieira¹⁷⁸  | Klaus von Gadow¹⁷⁹ |
 Hua-Feng Wang¹⁸⁰ | James Watson¹⁸¹ | Gijsbert D. A. Werner¹⁸² | Susan K. Wiser¹⁸³ |
 Florian Wittmann¹⁸⁴ | Verginia Wortel¹⁸⁵ | Roderick Zagt¹⁸⁶ | Tomasz Zawila-Niedzwiecki¹⁸⁷ |
 Chunyu Zhang¹⁸⁸ | Xiuhai Zhao¹⁸⁸  | Mo Zhou⁴ | Zhi-Xin Zhu¹⁸⁰ | Irie Casimir Zo-Bi⁸⁰ |
 Thomas W. Crowther¹

Correspondence

Iris Hordijk

Email: irishordijk@hotmail.com

Funding information

DOB Ecology

Swiss National Science Foundation, Grant/
Award Number: PZ00P3_193612VILLUM FONDEN, Grant/Award Number:
16549Instituto de Conservação da Natureza,
Grant/Award Number: UIDB/04033/2020

Handling Editor: Ellen Simms

Abstract

1. Biodiversity is an important component of natural ecosystems, with higher species richness often correlating with an increase in ecosystem productivity. Yet, this relationship varies substantially across environments, typically becoming less pronounced at high levels of species richness. However, species richness alone cannot reflect all important properties of a community, including community evenness, which may mediate the relationship between biodiversity and productivity. If the evenness of a community correlates negatively with richness across forests globally, then a

[Correction added on 10 May 2023, after first online publication: Author name Hyun Seok Kim has been corrected.]

greater number of species may not always increase overall diversity and productivity of the system. Theoretical work and local empirical studies have shown that the effect of evenness on ecosystem functioning may be especially strong at high richness levels, yet the consistency of this remains untested at a global scale.

2. Here, we used a dataset of forests from across the globe, which includes composition, biomass accumulation and net primary productivity, to explore whether productivity correlates with community evenness and richness in a way that evenness appears to buffer the effect of richness. Specifically, we evaluated whether low levels of evenness in speciose communities correlate with the attenuation of the richness–productivity relationship.

3. We found that tree species richness and evenness are negatively correlated across forests globally, with highly speciose forests typically comprising a few dominant and many rare species. Furthermore, we found that the correlation between diversity and productivity changes with evenness: at low richness, uneven communities are more productive, while at high richness, even communities are more productive.

4. *Synthesis.* Collectively, these results demonstrate that evenness is an integral component of the relationship between biodiversity and productivity, and that the attenuating effect of richness on forest productivity might be partly explained by low evenness in speciose communities. Productivity generally increases with species richness, until reduced evenness limits the overall increases in community diversity. Our research suggests that evenness is a fundamental component of biodiversity–ecosystem function relationships, and is of critical importance for guiding conservation and sustainable ecosystem management decisions.

KEYWORDS

diversity, ecosystem function and services, evenness, forests, global, productivity, species richness

1 | INTRODUCTION

In this era of diminishing biodiversity, understanding how changes in plant biodiversity will impact the functioning of ecosystems is critical (Ceballos et al., 2015; Isbell et al., 2017). Many studies show that the productivity of ecosystems typically increases with a greater number of species (Balvanera et al., 2006; Cardinale et al., 2007; Grace et al., 2016; Hooper et al., 2005). However, this positive relationship between productivity and diversity generally saturates and decays at high levels of species richness (Brun et al., 2019; Fei et al., 2018; Fraser et al., 2015; Liang et al., 2016; Schmid, 2002). Despite the consistency of this pattern across forests globally (Liang et al., 2016), we still lack a comprehensive understanding of the ecological relationships driving this attenuating effect of richness, which limits our capacity to relate diversity with productivity across the globe (Cardinale et al., 2012; Fraser et al., 2015).

The species redundancy hypothesis has been proposed to explain the saturating shape of the relationship between richness and ecosystem productivity (Cardinale et al., 2011; Loreau & Hector, 2001). This theory posits that, as the number of species in a region increases, functional redundancy increases and

the proportional impact of any single species on ecosystem functioning declines (Cardinale et al., 2011; Gitay et al., 1996). However, another possible explanation for the diminishing effect of diversity on productivity is that species richness only captures one aspect of community structure, and does not reflect the true ‘diversity’ of a community (Caswell, 1976; Stirling & Wilsey, 2001) (Figure 1). By reflecting the homogeneity of species abundances within the community, evenness serves as the other central component of diversity (Jost, 2010; Peet, 1974; Tuomisto, 2012). The relationship between community productivity and evenness might differ from the relationship between productivity and richness.

Broad-scale analyses suggest that speciose communities tend to be dominated by a few species, with a long tail of rare species (i.e. are highly uneven) (ter Steege et al., 2013). If such a negative relationship between richness and evenness holds across forests globally, then (i) speciose communities might tend to be dominated by a small number of species that have a disproportionate influence on ecosystem functioning, and (ii) the increasing richness of speciose communities might come at the expense of evenness. This trade-off might ultimately limit both functional diversity and ecosystem productivity.

Of course, a negative relationship between richness and evenness would only be relevant if both components of diversity influence the productivity of the community. Experimental studies provide clear evidence that community evenness has a direct effect on productivity (Kirwan et al., 2007; Sonkoly et al., 2019; Yan et al., 2021). Furthermore, if uneven communities tend to be dominated by one or a few species, the mass-ratio hypothesis suggests that increasing the number of rare species should have a relatively minimal impact on productivity (Grime, 1998; Lembrechts et al., 2018; Loreau, 1998). In contrast, in highly even communities, introducing a new species that is relatively abundant could substantially alter overall community productivity, for example through niche complementarity (see Figure 1) (Lembrechts et al., 2018; Nijs & Roy, 2000; Niklaus et al., 2017). Still, it remains unclear whether these effects of evenness mediate the effects of richness on productivity within complex natural communities and across biomes, because we do not know whether (i) the apparent negative relationship between richness and evenness is

consistent across forests globally (Soininen et al., 2012; Zhang, John, et al., 2012) and (ii) there is an interactive effect of evenness and richness on productivity across a range of environmental conditions. Disentangling these relationships between richness, evenness and productivity is critical for understanding the mechanisms governing the diversity–productivity relationship across broad spatial scales (Figure 1).

Here, we used approximately 1 million forest inventory plots (sourced from the Global Forest Biodiversity Initiative database) to explore whether community evenness affects the relationship between richness and productivity across global forests. If evenness restricts the positive effect of richness on ecosystem productivity, then two things must hold: (i) richness trades off with evenness across forests globally, such that speciose communities tend to be dominated by a few species and (ii) there is an interactive effect of both richness and evenness on the productivity of forest ecosystems. Specifically, we tested the hypothesis that (i) high species richness generally comes at the expense of evenness, as the greater number of species within

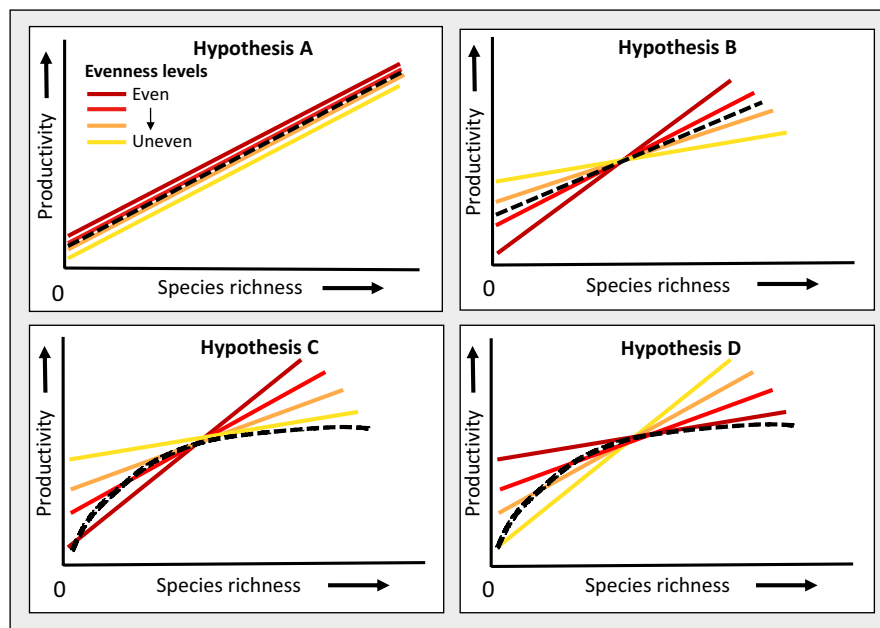


FIGURE 1 Four hypotheses describing how evenness interacts with the relationship between richness and productivity. The four levels of evenness are indicated with different colours from low evenness (yellow) to high evenness (red), and lower colour intensity indicates a lower data density. The dashed black line indicates the average relationship across the system. In all hypotheses, it is assumed that species richness has a positive effect on productivity (Balvanera et al., 2006; Cardinale et al., 2007; Hooper et al., 2005). If community evenness interacts with the effect of species richness on productivity, then (1) the effect of species richness on productivity depends on the evenness of the community, and (2) richness is correlated with evenness across the system (either positively or negatively). If there is no interaction between richness and evenness (Hypothesis A) or if there is no correlation between evenness and richness (Hypothesis B) (Ma, 2005), then the average effect (dashed line) of richness on productivity will neither attenuate nor increase at high richness levels. In such instances, the observed decrease in productivity at high richness levels is more likely a by-product of other ecological processes (e.g. functional redundancy). If, however, there is a significant interaction between richness and evenness, such that uneven communities have lower productivity at high richness (Hypothesis C), then a negative correlation between richness and evenness (Cook & Graham, 1996; Hanlin et al., 2000; Symonds & Johnson, 2008) would lead to an attenuation in productivity at high richness level: the marginal trend (dashed line) first tracks the high evenness isocline at low richness levels but then bends down towards the low evenness isoclines at high richness. Conversely, if uneven communities exhibit higher productivity at high richness levels (Hypothesis D), then a positive correlation between richness and evenness (Cotgreave & Harvey, 1994; Manier & Hobbs, 2006; Tramer, 1969) would explain this reduction in productivity: the marginal trend first tracks the low evenness isocline at low richness levels but then bends down towards the high evenness isoclines at high richness.

a community corresponds to a greater proportion of locally rare species, and (ii) that richness positively correlates with productivity in highly even, species-poor communities, but at the highest levels of species richness this positive relationship will break down due to an intrinsic reduction in evenness (Figure 1, Hyp. C).

2 | MATERIALS AND METHODS

To evaluate the relationship between evenness and richness, we incorporated all 1,011,027 forest inventory plots from the Global Forest Biodiversity Initiative (GFBI database, 2021, including data from Condit et al., 2019a, 2019b). Forest plot size ranged from 0.0008 to 2.0 ha. Because evenness and richness inherently vary with scale (Gleason, 1922; Wilson et al., 1999), plot sizes smaller than 0.02 and larger than 1.5 ha were excluded to ensure comparable results. Indeed, the effect of plot size on richness is especially strong when plots are smaller than 0.02 ha ($r=0.39$, $r^2=0.15$, $p<0.001$) or larger than 1.5 ha ($r=-0.84$, $r^2=0.71$, $p<0.001$). Elimination of these two plot size groups resulted in a filtered dataset with weak correlations between plot size and richness ($r=0.13$, $r^2=0.02$, $p<0.001$), and plot

size and evenness ($r=-0.09$, $r^2=0.008$, $p<0.001$). Although there is a range in plot sizes, 75% of the plots have a size between 0.02 and 0.06 ha. Quality controls of tree density estimations were conducted, and we removed plots with tree densities that fell outside the median ± 2.5 times the median absolute deviation, a moderately conservative threshold, within each biome (0.8% of total plots) (Leys et al., 2013). Additionally, we removed plots with unlikely biomass accumulation and productivity values (for details, see Section 2.2). The final dataset (Figure 2) comprised 896,276 forest plots, containing information on tree species richness and abundance. Of these, 367,565 plots contained diameter at breast height (DBH) information of individual trees which allowed us to estimate biomass. For both datasets, the plots were measured once between 1980 and 2017, with 2002 as the mean measurement year. The mean age of the forest where the plots are established is 52 years, with a standard deviation of 15 years (estimated with the global forest age map; Poulter et al., 2019).

Richness and evenness were calculated for each plot, enabling us to evaluate the nature of the relationship between these two components of diversity at biome and global scales. To examine the effect of evenness and richness on biomass accumulation and

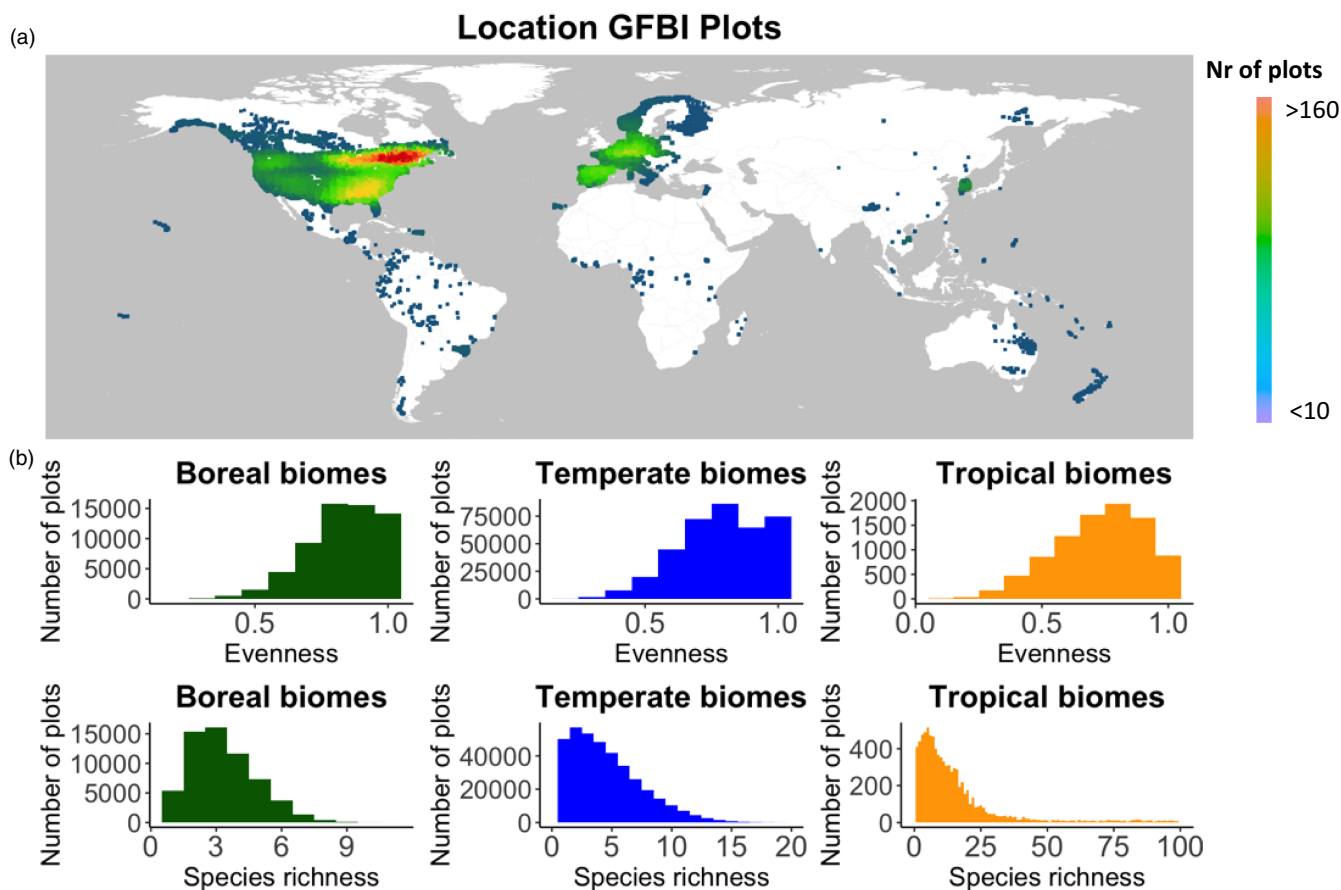


FIGURE 2 (a) Location of the Global Forest Biodiversity Initiative (GFBI) plots used in this study, where the density of forest plots is indicated from low density (blue) to high density (red). (b) The distribution of evenness and richness in the boreal, temperate and tropical biomes. An evenness value of one resembles either a monospecific stand or an even abundance of species. The tail of richness values of the tropical biomes extends to 380 species (not shown in the graph). The majority of our dataset is composed of secondary forests (mean age is 52 years), and especially the monospecific and relatively species-poor stands were affected by human activity in some degree.

productivity, we fit linear models that simultaneously accounted for evenness and richness while controlling for a wide range of environmental variables, including temperature, precipitation and soil characteristics.

We estimated forest productivity using two complementary approaches. First, we used biomass estimates, calculated using regional allometric equations that included information about DBH and species information from each individual tree, and divided this information by forest age (Poulter et al., 2019) to approximate stand-level biomass accumulation over time. Second, we estimated net primary productivity (NPP) for each plot using MODIS satellite data (Running et al., 2011), thus providing both ground-sourced and remotely sensed estimates of productivity. Satellite-derived NPP estimates largely agreed with ground measured temporal data for the plots where this information was available ($r=0.65$, $p<0.001$), see Section 2.3 for more details.

2.1 | Examining the relationship between species richness and evenness

Richness was defined as the total number of tree species per forest plot. We estimated Hill's evenness at the same plot level using the following equation:

$$\text{Evenness} = \exp(H)/S, \quad (1)$$

where H is Shannon's entropy, S is the species richness and $\exp(H)$ can be figuratively interpreted as an approximation of the number of relatively abundant species (Hill, 1973; Jost, 2010). Therefore, Hill's evenness can roughly be seen as the proportion of species that dominate the community in terms of abundance. Evenness values range from close to zero, when the community is dominated by only a few species, to one, where all species in the community have the same number of individuals.

There are many evenness indices available in the ecological literature (Tuomisto, 2012), and some commonly used indices are mathematically restricted by richness (including Pielou's J), reflect particularly dominance in the community (e.g. Simpson's evenness) or express evenness as the ratio between diversity and richness (including Hill's evenness). We also included the standardized Hill's evenness index in our supplementary analyses, as Hill's evenness reports an evenness of one in monospecific sites, which is avoided when using standardized Hill's evenness. As we explored the relationship between richness and evenness, each commonly used and distinct evenness index evaluated (standardized Hill's evenness, Pielou's J , Simpson's evenness, Evar, PIE, Eq) revealed similar overall trends (Figure S2) (Tuomisto, 2012). Given this overall consistency across different indices, including standardized Hill's evenness, we focus on the well-known Hill's evenness for the majority of our analyses because (i) it reflects evenness according to the current understanding of evenness as the ratio between diversity and richness (Tuomisto, 2012) and therefore (ii) consists

of the mathematically robust components of Shannon's entropy and richness (Hill, 1973).

In addition to analysing the relationship between evenness and richness at a global scale and for the major forest biomes (boreal, temperate and tropical biomes), for every World Wildlife Fund (WWF) biome we selected a number of plots that is proportional to the forested area within that biome (Table S2; Olson et al., 2001). By taking this subset of the data, we reduced the sampling bias by assuring that the results are more representative for either the forests globally or the boreal, temperate or tropical biomes. Regarding the global dataset, the majority of the plots fall within the temperate broadleaf and mixed forest biome and a smaller proportion of the data represents the tropical moist forest, but with subsetting the data we create a more representative global dataset. The same procedure is followed when subsetting biomes to represent the boreal, temperate and tropical biomes. The proportions of forested area for every biome were calculated in Google Earth Engine by overlaying the WWF biomes with a global map of existing forest cover (Hansen et al., 2013). Areas with more than 10% canopy closure for vegetation taller than 5m were defined as forests (FAO, 2000). To avoid heteroscedasticity due to the skewed nature of the species richness distribution, richness was log transformed, and the relationship was evaluated with a Pearson correlation (see Figure 3). To show that the relationship between evenness and richness is neither a mathematical artefact of the Hill's index nor dependent on the evenness index used, we visualized the results of a null model (Figure S1) and the relationship for commonly used evenness indices (Figure S2). The null model was created by evaluating the relationship between evenness and richness of 10,000 random data subsets, formed by drawing species according to a multinomial distribution from the dataset (see Figure S1 for a more detailed explanation). Additionally, to evaluate the effect of monospecific stands on the relationship between evenness and richness in forests globally, we compared the correlation coefficients of the relationship with and without monospecific stands.

2.2 | Biomass estimation

To estimate the above-ground biomass of each tree in extratropical biomes, we used 430 species-specific DBH-based allometric equations obtained from the GlobAllomeTree database to estimate the above-ground biomass of each tree (Henry et al., 2013). These allometric equations use a common logarithmic equation for estimating above-ground biomass from DBH measures (Jenkins et al., 2003):

$$\text{Biomass} = e^{(\beta_0 + \beta_1 \times \ln^{\text{DBH}})}, \quad (2)$$

where biomass is the total above-ground biomass (kg dry weight), DBH is the measured diameter at breast height (cm), \ln is the logarithm to the base e (2.718), and β_0 and β_1 are free parameters governing the effect of DBH on above-ground biomass. Following Jenkins et al. (2003), we applied back calculation to generate a pseudo dataset for biomass

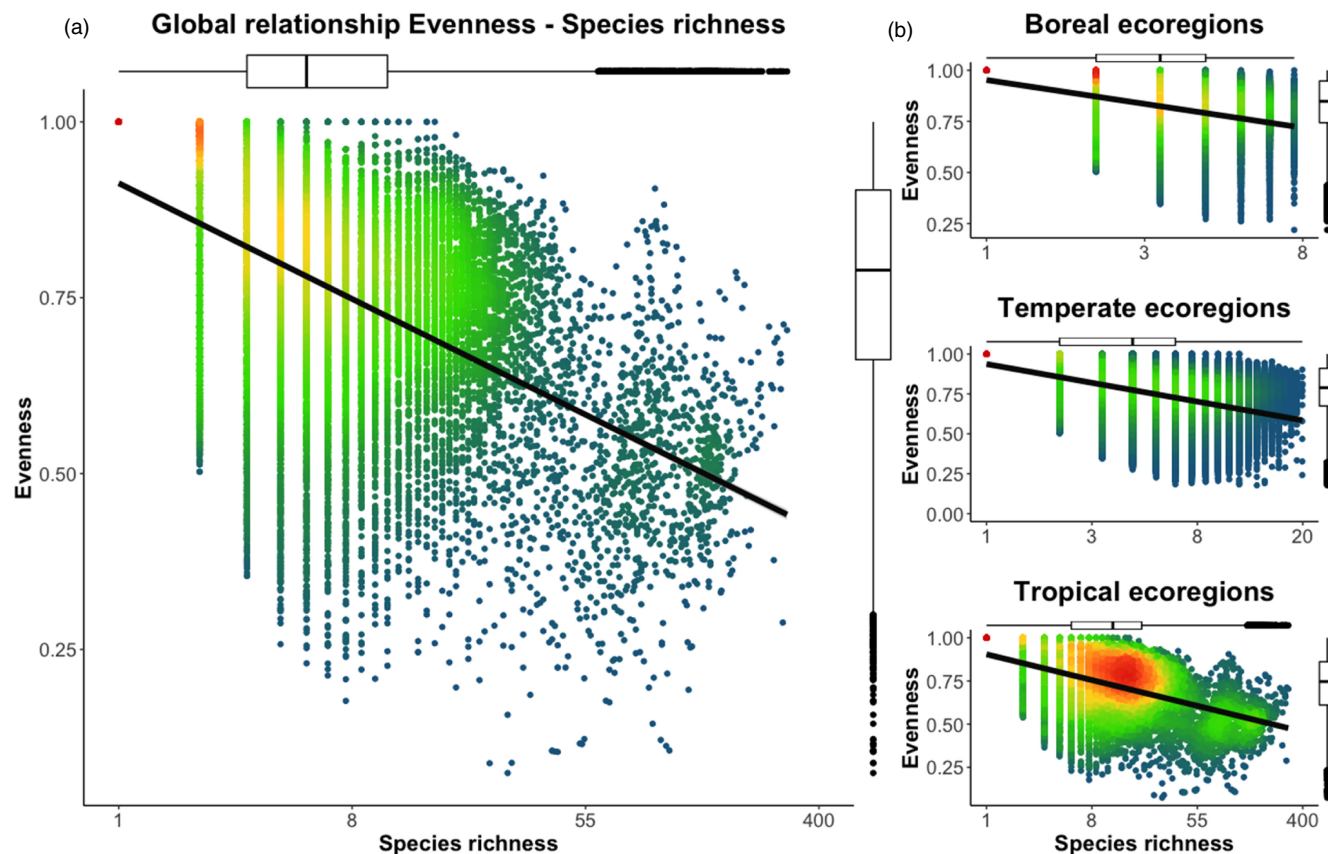


FIGURE 3 The relationship between evenness and logged species richness, where the data density is indicated from low density (blue) to high density (red). The Pearson correlation is highly significant for all relationships, we therefore describe the adjusted r^2 as a measure of effect size. (a) The global relationship between evenness and richness ($N=20,272$, $r=-0.52$, $r^2=0.28$). (b) The relationship between evenness and richness for the boreal ($N=61,712$, $r=-0.42$, $r^2=0.18$), temperate ($N=374,142$, $r=-0.53$, $r^2=0.28$), and tropical ($N=9570$, $r=-0.48$, $r^2=0.23$) biomes.

changes along DBH gradients based on each of the 430 allometric equations. To generate the pseudo data, we applied the following rules: (1) for a DBH between 5 and 25 cm, each centimetre was assigned a corresponding pseudo biomass value; (2) for a DBH between 25 and 100 cm, every 5 cm were assigned a corresponding value; (3) for a DBH between 100 and 300 cm (maximum DBH), every 10 cm were assigned a corresponding value. Consequently, we trained biome-specific allometric equations for each biome (varying in the β_0 and β_1 parameter estimates) based on the pseudo-DBH and biomass dataset (Olson et al., 2001) (Table S1; Figure S3).

Biomass estimations for the tropical biomes followed the allometric equations for pantropical regions from Chave et al. (2014), which are available through the R package 'BIOMASS' (Réjou-Méchain et al., 2017). These equations require information on wood density, and we compiled species-specific wood density estimates from the Global Wood Density Database (Chave et al., 2009) and the BAAD database (Falster et al., 2015). To match the binomial species names between the GFBI and the wood density databases, we standardized species binomials using the Taxonomic Name Resolution Service platform (Boyle et al., 2013).

After computing the above-ground biomass for all 24 million individuals in our dataset, plot-level biomass values were obtained

by summing up the biomass of all individuals in the respective plot. Biomass densities (t/ha) of each plot were obtained by dividing the total above-ground biomass (t) by the plot size (ha).

2.3 | Biomass accumulation rates and productivity data

We selected approximately 95% of the data by excluding plots that had biomass values greater than two times the standard deviation above and below the mean plot-level biomass per biome to filter potential outliers in biomass values due to errors in measurement or data management. Additionally, we excluded plots with biomass values higher than 1 million kg/ha, as these values were likely overestimating biomass due to the presence of a large tree in the plot (Bastin et al., 2018; Chave et al., 2004; Slik et al., 2013). This filtering procedure resulted in 367,565 plots that were used in the final analyses (see Table S2 for an overview of the number of plots for the biomass and productivity analyses). All biomass values were divided by estimated forest age (Poulter et al., 2019), to control for differences in forest developmental stage that cause striking differences in forest biomass accumulation (Peichl & Arain, 2006; Poorter et al., 2016).

As such, final biomass estimates reflect the mean annual biomass accumulation over time.

Biomass accumulation over time is not a precise indicator of forest productivity, as it cannot account for all of the ecological dynamics and disturbances that occurred during forest development, but it is a useful proxy for overall cumulative growth at the ecosystem level. However, to attempt to account for the annual variation in forest productivity, we used productivity estimates from an independent satellite-derived product. Specifically, we estimated NPP for every plot location from MODIS satellite images, using Google Earth Engine (Gorelick et al., 2017; Running et al., 2011) to supplement the forest biomass analysis. NPP is calculated as absorbed fraction of photosynthetically active radiation, which is a combination of leaf area index and fraction of photosynthetically active radiation, while taking temperature and water stress into account as well (Running et al., 2011). We calculated the mean productivity between 2000 (first year of NPP data availability) and 2009 (third quartile of forest age in our data), to obtain a robust NPP value and coincide the average NPP measurement year with the median forest age in our dataset. In the temperate biomes—the only region where sufficient DBH information was available across multiple years in the GFBI database—the ground-measured temporal changes in biomass were fairly well correlated with the satellite-derived productivity data ($r=0.65$, $p<0.001$). The lack of forest plots measured multiple times in the boreal and tropical biomes, precluded the estimation of the accuracy of the NPP data from these regions. We conducted all analyses across both mean annual biomass accumulation and NPP to explore the unifying trends that emerge across both approaches.

2.4 | Evaluating the effect of evenness and richness on biomass accumulation and productivity

We used linear models to assess the effect of evenness, richness and their interaction, on NPP and biomass accumulation. The data met the linear model assumptions and evaluation of the model fit (Q-Q plots, distribution of residuals) suggested that linear models were applicable. To account for the fact that biomass accumulation does not necessarily vary linearly with time or successional state, we included time as a covariate in the model of biomass accumulation rate, which is equivalent to allowing the relationship between stand-level biomass and time to be quadratic (Poulter et al., 2019). Additionally, we controlled for the potentially confounding effects of climatic drivers and other environmental influences (Ali et al., 2019), including mean annual temperature, isothermality, annual precipitation, variation in seasonal precipitation (Hijmans et al., 2005), and soil carbon, sand content and pH in the upper 15 cm of the soil as covariates, each extracted from global maps at a 30-arc second resolution (Batjes et al., 2017; Ribeiro et al., 2018). These climate variables capture both the mean and variation in temperature and

rainfall, while the edaphic variables include the most important soil drivers of biomass and productivity (Ali et al., 2019). Human impact was also considered in the linear models by including estimates of the percentage of human development, calculated as the percentage urban areas and managed vegetation per square kilometre (Tuanmu & Jetz, 2014), and population density as the number of people per square kilometre (Center for International Earth Science Information Network—CIESIN—Columbia University, 2016). The density of trees was included as an independent variable in the analyses, as this can affect both evenness (Wilson et al., 1999) and richness values (Lomolino, 2000). Additionally, 'biome' was included as a factor in the global model, and we accounted for plot size in all models (Poulter et al., 2019). To explore the possible effect of plot size, we analysed the relationship between evenness and richness and the effect of both on the two measures of productivity for small, medium and large plot sizes for the dataset globally (Figure S5). Additionally, we performed a sensitivity analysis on species richness, evenness and their interaction to plot size when predicting NPP or biomass accumulation rates in the global dataset (Figure S6A) and on biome level (Figure S6B–D).

Collinearity between variables in our model was evaluated using variance inflation factors (VIF). Although evenness and richness were negatively correlated, on average, there was substantial variation in this trend (Figure 3), such that the VIF values between evenness and richness were less than 3.0 in all models, and both were therefore included in every model (Becker et al., 2015). The model was implemented at the global scale, in which every biome was proportionally represented according to the extent of forested area within that biome (as described above), and at the biome level. At the biome level, the boreal, temperate and tropical biomes were represented by their largest forested ecosystems, respectively, the boreal forest, temperate broadleaf and mixed forest, and tropical moist forest. To visualize how richness and evenness interacted to affect productivity, we plotted the relationship between richness and productivity at four different levels of evenness (<0.4, 0.4–0.6, 0.6–0.8, and 0.8–1). The lowest evenness level <0.4 was chosen as this represents the lower end of the evenness values in the dataset, while including sufficient observations in this category. The subsequent categories were defined as an increase of 20% in the evenness values. For each evenness level, productivity was predicted with every model covariate (e.g. climate, soil, human impact) set at its median value. Thus, each predicted productivity is marginal to the aggregate trend at the global scale and for each biome. To examine whether the relationship between richness, evenness and productivity varied with evenness index, we evaluated the global trend for six commonly used evenness indices (Figure S7). We used a bootstrapping approach to incorporate uncertainty in the biomass calculations and satellite-derived NPP (Figure S8).

To quantify the relative importance of evenness, richness, their interaction, environmental variables and human impact, we used the scaled calc.relimp function in R (Grömping, 2006). This function evaluates the contribution of each independent variable to the

variation explained, by averaging the contribution of each independent variable to the r^2 in terms of its sum of squares across all possible fitting sequences. The statistical analyses were performed in R version 3.5.1 (R Core Team, 2018). The code used to perform the statistical analyses can be found at Github, following this link: [tinyurl.com/3vfvf52v9](https://github.com/3vfvf52v9).

3 | RESULTS

There was a consistent, negative correlation between evenness and richness at the global scale, which was apparent across all forest biomes (Figure 3a–d), scales (Figure S4) and plot sizes (Figure S5). This negative correlation was also found using several other commonly used evenness indices, including the normalized Hill's evenness index, Simpson's evenness, E_{var} index and PIE (Figure S2), suggesting that the relationship is robust to the evenness index used. Moreover, the magnitude of the negative correlation was more negative than expected at random, suggesting that the relationship between evenness and richness is not a mathematical artefact of the evenness metrics used (Figure S1). Instead, it demonstrates that highly speciose, highly even forest communities are less rare in nature than would be expected by chance (Figure S1), suggesting that biotic processes may play a role in shaping the relationship between richness and evenness. Additionally, excluding monospecific stands did not change the sign of the correlation, although the relationship was stronger when monospecific stands were included as they force the correlation through one ($r = -0.08$ for forests globally, $r = -0.07$ excluding monospecific stands). Within forests, species-poor communities tend to have a relatively uniform distribution of abundances, while among speciose communities, having more species is associated with increasingly uneven abundance distributions caused by a few dominant and many rare species (Figure 3a–d). Temperate and boreal biomes exhibited a saturating relationship, with evenness never extending below $E = 0.75$, on average, even at the highest richness levels.

However, after accounting for underlying environmental variation, we identified a significant positive relationship between richness and both biomass accumulation and productivity ($0.06 < \beta < 0.48$, $p < 0.01$), supporting a recent global analysis (Liang et al., 2016) (but see also (Sheil & Bongers, 2020)). However, the relationship varied among regions, as temperate forests exhibited a marginally negative relationship between richness and productivity ($\beta = -0.09$, $p < 0.05$) (Figure 4). Given the observed negative relationship between richness and evenness, increasing evenness was generally associated with lower biomass accumulation. However, this pattern was not observed in tropical forests ($\beta = 0.04$, $p < 0.01$), where there was a positive relationship between evenness and productivity ($0.004 < \beta < 0.014$, $p < 0.01$) (Figure 4).

In combination, both the evenness and richness of forest communities were more strongly related to plot-level biomass accumulation than to productivity (25% vs. 19% of variance explained in the global model) (Figure 4). Among boreal forests, which are species

poor, richness was more strongly related to both biomass accumulation and productivity than was evenness (variance explained by richness_{Biomass} = 10.1, evenness_{Biomass} = 6.8 and richness_{NPP} = 14.8, evenness_{NPP} = 0.4) (Figure 5b,f). In contrast, among moist tropical forests, which are species rich, productivity was similarly related to either richness or evenness (variance explained by richness_{NPP} = 1.8, evenness_{NPP} = 1.6) (Figure 5d,h).

Experimentally increasing species richness increases biomass accumulation (Balvanera et al., 2006; Cardinale et al., 2007; Grace et al., 2016; Hooper et al., 2005), but observational data suggest that the effect becomes weaker or even reverses at high species richness (Brun et al., 2019; Fei et al., 2018; Fraser et al., 2015; Liang et al., 2016; Schmid, 2002). We hypothesized that if increasing community evenness causes productivity to decline, then the attenuating effect of richness on productivity could be explained by a negative correlation between species richness and community evenness (Figure 1). To explore if productivity data from the world's forests might be explained by such a hypothesis, we modelled the data with a regression featuring the hypothesized interaction between richness and evenness (Figure 5). The proposed interaction between richness and evenness is statistically significant in every model ($p < 0.05$), although the strength of the interaction varies. The strongest interactive effect was observed on biomass accumulation in the moist tropical forest, which are highly speciose ($\beta = 0.14$) (Figure 5d). A weaker interaction was observed in temperate and boreal forests ($\beta = 0.03$ and $\beta = -0.06$, respectively), which contained fewer species (Figure 5b,c). In the tropics, biomass accumulation is predicted to vary by 2.5-fold between high- and low-evenness communities (6000 vs. 15,000 kg/ha year⁻¹ at maximum richness) (Figure 5c), whereas this range is much lower in the temperate (4700–5000 kg/ha year⁻¹ at maximum richness) (Figure 5c) and boreal systems (850–1150 kg/ha year⁻¹ at maximum richness) (Figure 5b). Moreover, we find the same trend for productivity (Figure 5f–h), across different evenness indices (Figure S7), plot sizes (Figure S5) and when incorporating uncertainty of the biomass and productivity values (Figure S8). Additionally, the sensitivity analyses show that there is no consistent effect or no effect of plot size on the relationship of evenness, richness and their interaction on both measures of productivity (Figure S6).

4 | DISCUSSION

Among forests across the globe, the positive correlation between species richness and ecosystem productivity appears to attenuate in the most speciose communities (Liang et al., 2016). Our analysis examines whether the evenness of plant communities might contribute to this attenuation. We observed a consistent negative correlation between richness and evenness across forests, globally (Figure 3), whereby highly speciose communities exhibited relatively low levels of evenness. We also detected interactive effects of the correlation of richness and evenness with ecosystem productivity, which lends support to the hypothesis that evenness might mediate the impacts of richness on

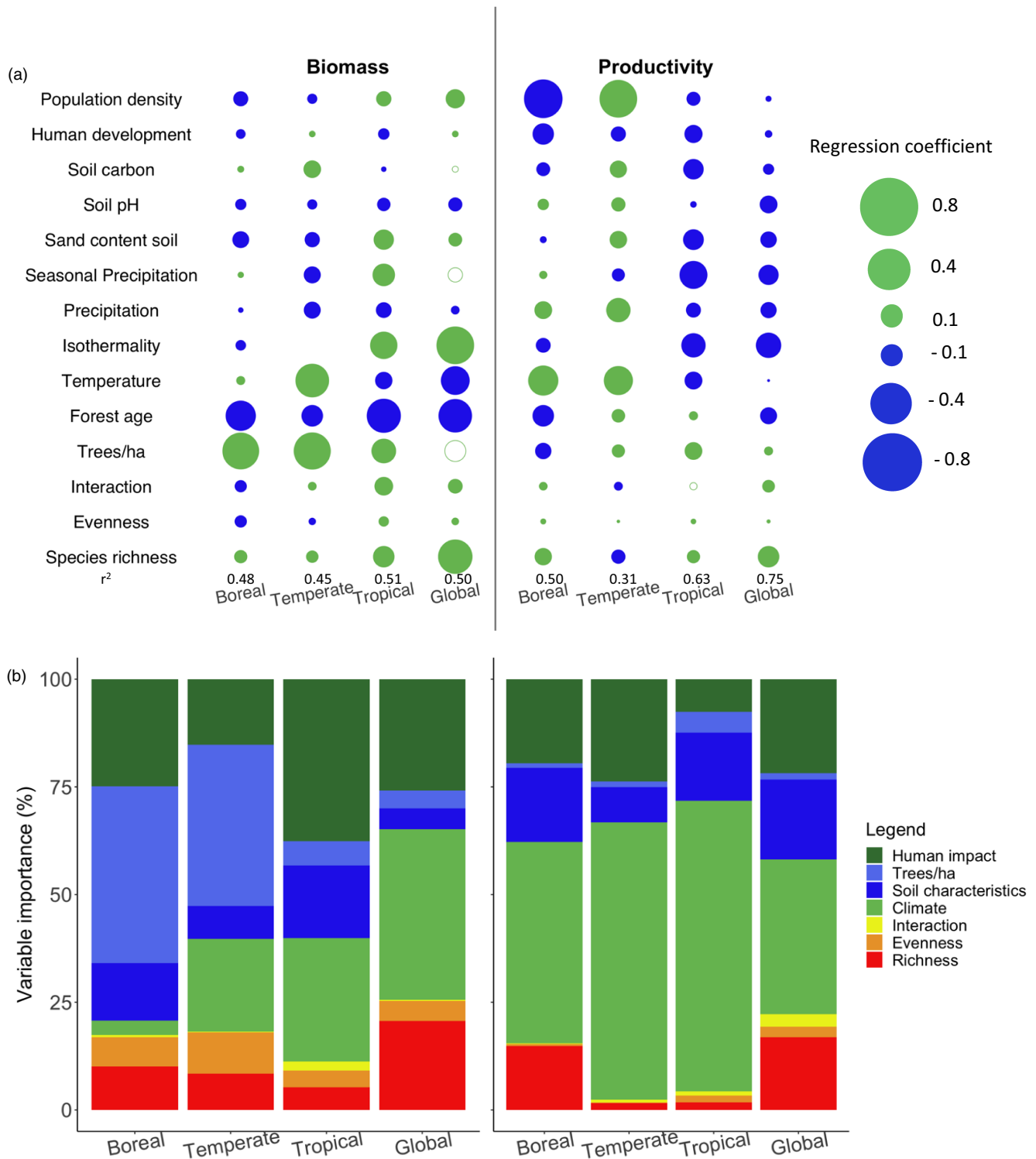


FIGURE 4 (a) Positive (green) and negative (blue) regression coefficients, and (b) variable importance of evenness, richness, the interaction of evenness and richness and climate, soil and human impact variables on biomass and productivity. Only the results for the boreal, temperate, tropical moist forest and all the biomes globally are visualized, and variables causing multicollinearity are taken out (see Section 2). The open circles in (a) indicate non-significant coefficients, while the filled circles indicate significant coefficients. The adjusted r^2 values of the linear models are displayed in (a).

forest productivity. In short, when the number of species is relatively low, species richness correlates positively with productivity. However, in the most speciose communities, the long tail of rare species makes

them highly uneven, which may limit the impact of diversity on productivity. As such, in these highly uneven communities, a greater number of species does not necessarily correlate with greater productivity.

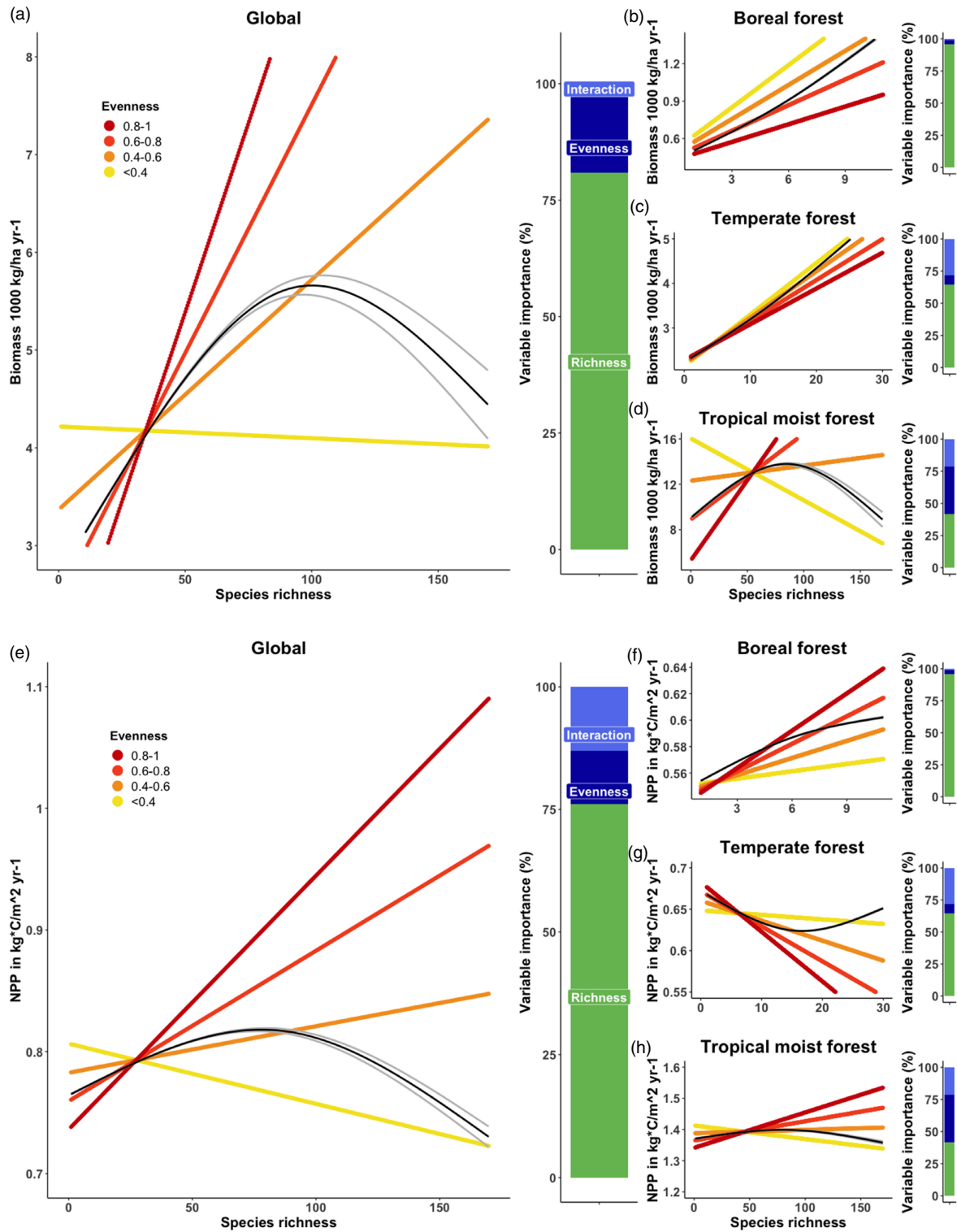


FIGURE 5 The hypothesized effect of different levels of evenness on the relationship between species richness and mean annual biomass accumulation (a–d) or productivity (h–e). The graphs visualize predicted values based on the results of a linear model, with the covariates held constant (see methods), and as a cut-off point the third quantile of the biomass and NPP values to avoid overfitting. The data are projected on the graph (black line), and the 95% upper and lower confidence intervals are visualized in grey. At the right side of every graph the scaled variable importance, according to a linear model including covariates, of richness (green), evenness (dark blue), and the interaction between richness and evenness (light blue) is visualized. In the graphs at the left side, the global effect is visualized, while at the right side the data are split among boreal, temperate and moist tropical forests. The uncertainty of the biomass calculations and estimated productivity are visualized in [Figure S7](#).

Our results suggest that, when there are more species present, they are more likely to have an uneven distribution. However, this relationship between richness and evenness was significantly more negative for our observational data than would be expected under null-model predictions (Figure S1), highlighting this pattern is at least partially a by-product of ecological and evolutionary processes. The observed negative relationship between evenness and richness observed across forests globally (Figure 3) may potentially arise if species-rich ecosystems contain relatively fewer dominant, but more rare species than species-poor systems. We focus on the Hill's evenness index because it is largely uncorrelated with richness (Hill, 1973), to ensure that the observed negative correlation between richness and evenness is not merely a mathematical artefact of the evenness index used (Figure S1). Nevertheless, our results are robust to a range of common evenness indices (Figures S2 and S7), demonstrating the generality of this relationship and of the resulting global trends. The drivers of this trend are likely to include a range of ecological coexistence mechanisms. A possible explanation for this negative correlation could be that in cold boreal forests, there are only a few species present, and intraspecific competition is expected to be higher than the interspecific competition (Aguiar et al., 2001). However, in warm, moist environments, such as moist tropical forests, there are many species present. Higher species richness is likely to give rise to substantial interspecific competition, as well as scope for additional ecological mechanisms such as the Janzen Connell effect to influence the abundance of species (Connell, 1971; Janzen, 1970). As species have different competitive abilities and strategies, high levels of asymmetric competition allow relatively few species to become dominant, with the majority of the species being either abundant or rare (McGill et al., 2007).

The strength of the negative correlation between evenness and productivity varies between biomes (Figure 5) (Sokoly et al., 2019; Zhang, Chen, & Reich, 2012), which may be indicative of a greater effect of niche partitioning in more speciose and even forests. In biomes with few species, increasing the richness of species tends to enhance resource partitioning and productivity (Figure 5b,c) (Isbell et al., 2009; van Ruijven & Berendse, 2005). In contrast, in biomes with many species, evenness of those species is relatively low as there are many rare species, and so increasing the evenness might increase ecosystem productivity (Kirwan et al., 2007; Zhang, John, et al., 2012; Zhang, Chen, & Reich, 2012) (Figure 5d,h). Interestingly, a different trend was observed for productivity in temperate forests, where at high evenness productivity peaks at low richness, and at low evenness productivity peaks at high richness (Figure 5c). This could possibly be explained by the overarching importance of environmental drivers for productivity in this biome, with the combined effects of richness and evenness explaining only 2% of the total explained variance (Figure 4b). Conversely, this biome contains the greatest proportion of forest plots with high human activity (Figure 4b), such that this trend may reflect different management practices across biomes.

Our results suggest that it may not only be the redundancy of the species that drives the attenuating effect of the relationship

between richness and ecosystem productivity (see Cardinale et al., 2011), but also the low abundances (and corresponding minimal contribution to productivity) of the rare species that flattens the relationship between richness and productivity at high richness levels. Our hypothesis predicts that, within speciose communities, adding new species at extremely low abundances will have little impact on the overall productivity, relative to the effect on productivity of adding new species to less speciose communities. We stress that this finding does not discount the importance of rare species, which contribute significantly to productivity through positive complementarity effects in many ecological communities (Loreau & Hector, 2001; Sokoly et al., 2019). Indeed, rare species contribute to ecosystem multifunctionality, which can have very important indirect effects on productivity (e.g. being critical for pollination) (Dee et al., 2019; Lyons et al., 2005), or sustain productivity over the long term when they become more abundant with (environmental) change (Loreau & Hector, 2001; Loreau et al., 2003; Yachi & Loreau, 1999). However, our analysis supports the idea that—in line with the mass-ratio hypothesis—rare species tend to contribute less to productivity than do dominant species. If this hypothesis were correct, then a higher relative abundance of rarer species (i.e. increasing evenness) would increase the biodiversity and productivity within the system.

The two measures of productivity used here—biomass accumulation and satellite-derived NPP—each have their unique drawbacks and challenges (Sheil & Bongers, 2020), such as relying on coarse estimates of forest age, being susceptible to mismatches between plot size and satellite resolution, and uncertainty in the calculations. For the temperate region, satellite-derived NPP and productivity calculated from multiple times measured ground-sourced data were well correlated ($r=0.65$); however, we could not estimate the accuracy for the other biomes due to limited data. Despite the uncertainty in the biomass and productivity calculations, the main results were robust when considering the uncertainty within these two estimates of productivity individually (Figure S8). We chose to consider these metrics in tandem specifically to minimize the data limitations of each, and ensure that our results are qualitatively robust to the choice of productivity metric. Additionally, in our dataset, we have considerable variability in plot sizes, which can affect both species richness and evenness values, and subsequently the importance of niche and neutral processes (Gleason, 1922; Viana & Chase, 2019; Wilson et al., 1999). Yet, sensitivity analyses show that the differences in plot size do not change the main results (Figure S5), with no consistent biases across plot sizes (Figure S6). However, we do detect considerable noise across the range of plot sizes, partially due to unbalanced sample sizes in relation to plot size in between biomes.

Although our work establishes baseline empirical trends, it is important to highlight that these results rely on correlative trends using observational data from a compilation of different databases and sources. Future experimental research will be needed to test the hypotheses we have presented and to explore direct causal relationships among evenness, richness and productivity.

Long-term experimental studies with various plot sizes and successional stages will be critical for disentangling the relative importance of the different processes underpinning diversity and function. By being able to manipulate species richness, evenness and functional redundancy, while also obtaining direct temporal measurements of productivity, experimental studies will be critical for identifying mechanistic drivers that are difficult to assess from broad-scale observational and statistical approaches (Paquette et al., 2018). Nevertheless, our analysis is consistent with the findings from a wide range of biodiversity–ecosystem function experiments, observational and modelling studies, highlighting that the positive effect of species richness on forest productivity declined in the most speciose communities (Cardinale et al., 2007; Hooper et al., 2005). In addition, by exploring how these trends vary across biomes, this analysis can help to provide the context for the highly variable relationships between richness and productivity across the globe.

5 | CONCLUSIONS

An ever-growing body of evidence suggests that, across forests globally, plant species richness correlates with ecosystem productivity of plant communities (Liang et al., 2016; Luo et al., 2019). Our results support previous studies, showing that richness correlates positively with ecosystem productivity, particularly in communities with few species. However, as the number of species increases, the relationship between richness and productivity attenuates. We observed that, as species richness increases, the evenness of those communities tends to decline, which may potentially limit the influence of richness on ecosystem productivity in the most speciose communities. Because communities with many species tend to be dominated by relatively few species, the evenness of those communities tends to be relatively low. This negative correlation between richness and evenness may partially contribute to the attenuating effect of species richness on ecosystem productivity observed in highly diverse communities (Liang et al., 2016). In addition, this apparent trade-off between richness and evenness may explain some of the idiosyncrasies observed in previous biodiversity–productivity analyses, as the slope of the richness–productivity relationship will vary considerably across time and space due to the confounding effect of evenness. These trends have direct implications for ecosystem management practices by showing where community productivity is most dependent on richness or evenness. Ultimately, the interacting effects of richness and evenness help shape our understanding of the biodiversity–productivity relationship, identifying core relationships that link community structure to the functioning of forest ecosystems worldwide.

AUTHOR CONTRIBUTIONS

Iris Hordijk, Tom W. Crowther, Daniel S. Maynard and Simon P. Hart conceived of the study. Iris Hordijk extracted and analysed the data,

and drafted the manuscript with assistance from Dan Maynard and Tom Crowther. Biomass calculation was carried out by Lidong Mo. Members of the GFBI consortium (all authors not mentioned previously) provided data for the analysis. All authors assisted with revisions and gave final approval for publication.

AFFILIATIONS

- ¹Institute of Integrative Biology, ETH Zurich (Swiss Federal Institute of Technology), Universitätsstrasse 16, 8092, Zurich, Switzerland
- ²School of Biological Sciences, University of Queensland, Brisbane, Australia
- ³Naturalis Biodiversity Centre, Leiden, The Netherlands
- ⁴Systems Ecology, Free University Amsterdam, Amsterdam, The Netherlands
- ⁵Department of Forestry and Natural Resources, Purdue University, West Lafayette, Indiana, USA
- ⁶Department of Crop and Forest Sciences, University of Lleida, Lleida, Spain
- ⁷Joint Research Unit CTFC–AGROTECNIO–CERCA, Solsona, Spain
- ⁸Wageningen University and Research, Wageningen, The Netherlands
- ⁹Department of Forest Resources, University of Minnesota, St Paul, Minnesota, USA
- ¹⁰Hawkesbury Institute for the Environment, Western Sydney University, Penrith, New South Wales, Australia
- ¹¹Swiss Federal Institute for Forest, Snow and Landscape Research, WSL, Birmensdorf, Switzerland
- ¹²UFR Biosciences, University Félix Houphouët-Boigny, Abidjan, Côte d'Ivoire
- ¹³Faculty of Science and Technology, Free University of Bolzano, Bolzano, Italy
- ¹⁴Department of Agricultural, Food, Environmental and Animal Sciences, University of Udine, Udine, Italy
- ¹⁵Spatial Ecology and Conservation Laboratory, Center for Latin American Studies, University of Florida, Gainesville, Florida 32611, USA
- ¹⁶Forestry School, Tecnológico de Costa Rica TEC, Cartago, Costa Rica
- ¹⁷Fundacion ConVida, Universidad Nacional Abierta y a Distancia, UNAD, Medellín, Colombia
- ¹⁸Field Museum of Natural History, Chicago, Illinois, USA
- ¹⁹Center for Tropical Research, Institute of the Environment and Sustainability, UCLA, Los Angeles, California, USA
- ²⁰Silviculture and Forest Ecology of the Temperate Zones, University of Göttingen, Göttingen, Germany
- ²¹Division of Forest and Forest Resources, Norwegian Institute of Bioeconomy Research (NIBIO), Ås, Norway
- ²²Museo de Historia natural Noel Kempff Mercado, Santa Cruz, Bolivia
- ²³European Commission, Joint Research Centre, Ispra, Italy
- ²⁴UNELLEZ-Guanare, Programa de Ciencias del Agro y el Mar, Herbario Universitario (PORT), Portuguesa, Venezuela
- Compensation International S. A. Ci Progress-GreenLife, Bogotá, Distrito Capital, Colombia
- ²⁵School of Geography, University of Leeds, Leeds, UK
- ²⁶Department of Geomatics, Forest Research Institute, Raszyn, Poland
- ²⁷Centro Multidisciplinar, Universidade Federal do Acre, Rio Branco, Brazil
- ²⁸Proceedings of the National Academy of Sciences, Washington, DC, USA
- ²⁹Department of Evolutionary Anthropology, Duke University, Durham, North Carolina 27708, USA
- ³⁰Gembloux Agro Bio-Tech, University of Liege, Liege, Belgium
- ³¹United Nation Framework Convention on Climate Change, Bonn, Germany
- ³²Cirad, UMR-AMAP, CNRS, INRA, IRD, Université de Montpellier, Montpellier, France

- ³³Institute of Tropical Forest Conservation, Mbarara University of Sciences and Technology, Mbarara, Uganda
- ³⁴Isotope Bioscience Laboratory—ISOFYs, Ghent University, Ghent, Belgium
- ³⁵Integrated Center for Research, Development and Innovation in Advanced Materials, Nanotechnologies, and Distributed Systems for Fabrication and Control (MANSiD), Stefan cel Mare University of Suceava, Suceava, Romania
- ³⁶Department of Forest Sciences, Luiz de Queiroz College of Agriculture, University of São Paulo, Piracicaba, Brazil
- ³⁷Bavarian State Institute of Forestry, Freising, Germany
- ³⁸Spatial Ecology and Conservation Laboratory, School of Forest, Fisheries, and Geomatics Sciences, University of Florida, Gainesville, Florida 32611, USA
- ³⁹Institute of Biology, Geobotany and Botanical Garden, Martin Luther University Halle-Wittenberg, Halle-Wittenberg, Germany
- ⁴⁰German Centre for Integrative Biodiversity Research (iDiv) Halle-Jena-Leipzig, Leipzig, Germany
- ⁴¹Department of Agriculture, Food, Environment and Forest (DAGRI), University of Firenze, Florence, Italy
- ⁴²Biological Institute, Tomsk State University, Tomsk, Russia
- ⁴³Department of Spatial Regulation, GIS and Forest Policy, Institute of Forestry, Belgrade, Serbia
- ⁴⁴Department of Ecology and Evolutionary Biology, University of Connecticut, Storrs, Connecticut, USA
- ⁴⁵Tropical Forests and People Research Centre, University of the Sunshine Coast, Maroochydore, Queensland, Australia
- ⁴⁶Faculty of Natural Resources Management, Lakehead University, Thunder Bay, Ontario, Canada
- ⁴⁷IFER—Institute of Forest Ecosystem Research, Jilove u Prahy, Czech Republic
- ⁴⁸Global Change Research Institute CAS, Brno, Czech Republic
- ⁴⁹Nicholas School of the Environment, Duke University, Durham, North Carolina, USA
- ⁵⁰Department of Biology, University of Missouri-St Louis, St Louis, Missouri, USA
- ⁵¹Programa de Pós-graduação em Biologia Vegetal, Instituto de Biologia, Universidade Estadual de Campinas, Campinas, Brazil
- ⁵²Department of Plant Sciences and Conservation Research Institute, University of Cambridge, Cambridge, UK
- ⁵³Andes to Amazon Biodiversity Program, Madre de Dios, Peru
- ⁵⁴Facultad de Ciencias Forestales, Universidad Juárez del Estado de Durango, Durango, Mexico
- ⁵⁵Department of Physical and Biological Sciences, The College of Saint Rose, Albany, New York, USA
- ⁵⁶Department of Biology, West Virginia University, Morgantown, West Virginia, USA
- ⁵⁷Biology Department, Centre for Structural and Functional Genomics, Concordia University, Montreal, Quebec, Canada
- ⁵⁸Natural Science Department, Universidade Regional de Blumenau, Blumenau, Brazil
- ⁵⁹World Agroforestry (ICRAF), P.O. Box 30677 00100, Nairobi, Kenya
- ⁶⁰Cirad, UMR EcoFoG (AgroParistech, CNRS, INRAE, Université des Antilles, Université de la Guyane), Campus Agronomique, Kourou, French Guiana
- ⁶¹Department of Geographical Sciences, University of Maryland, College Park, Maryland, USA
- ⁶²Institute of Forestry, Belgrade, Serbia
- ⁶³National Institute of Amazonian Research, Manaus, Brazil
- ⁶⁴IRET, Herbar National du Gabon (CENAREST), Libreville, Gabon
- ⁶⁵Department of Ecology and Evolutionary Biology, University of Arizona, Tucson, Arizona, USA
- ⁶⁶Queensland Herbarium, Department of Environment and Science, Toowong, Queensland, Australia
- ⁶⁷Ecole de Foresterie et Ingénierie du Bois, Université Nationale d'Agriculture, Ketou, Benin
- ⁶⁸Biology Centre of the Czech Academy of Sciences, Institute of Entomology, Ceske Budejovice, Czech Republic and Institute for Tropical Biology and Conservation, Universiti Malaysia Sabah, Kota Kinabalu, Sabah, Malaysia
- ⁶⁹Museu Paraense Emílio Goeldi. Coordenação de Ciências da Terra e Ecologia, Belém, Pará, Brazil
- ⁷⁰Geography, College of Life and Environmental Sciences, University of Exeter, Exeter, UK
- ⁷¹Natural Resources Institute Finland (Luke), Joensuu, Finland
- ⁷²Institute of Plant Sciences, University of Bern, Bern, Switzerland
- ⁷³Forest Research Institute Malaysia, Kuala Lumpur, Malaysia
- ⁷⁴Department of Sustainable Agro-Ecosystems and Bioresources, Research and Innovation Center, Fondazione Edmund Mach, San Michele all'Adige, Italy
- ⁷⁵Forestry Division, Food and Agriculture Organization of the United Nations, Rome, Italy
- ⁷⁶School of Forestry and Environmental Studies, Yale University, New Haven, Connecticut, USA
- ⁷⁷Royal Botanic Garden Edinburgh, Edinburgh, UK
- ⁷⁸Department of Plant Sciences, University of Oxford, Oxford, UK
- ⁷⁹Department of Plant Systematics, University of Bayreuth, Bayreuth, Germany
- ⁸⁰Cirad, UPR Forêts et Sociétés, University of Montpellier, Montpellier, France
- ⁸¹Department of Forestry and Environment, National Polytechnic Institute (INP-HB), Yamoussoukro, Côte d'Ivoire
- ⁸²Centre for Conservation Science, The Royal Society for the Protection of Birds, Sandy, UK
- ⁸³Wild Chimpanzee Foundation, Liberia Office, Monrovia, Liberia
- ⁸⁴Instituto de Investigaciones de la Amazonía Peruana, Iquitos, Peru
- ⁸⁵Department of Mathematical Sciences, Centre for Invasion Biology, Stellenbosch University, Stellenbosch, South Africa
- ⁸⁶Theoretical Ecology Unit, African Institute for Mathematical Sciences, Cape Town, South Africa
- ⁸⁷Division of Forest Resources Information, Korea Forest Promotion Institute, Seoul, South Korea
- ⁸⁸Institut Agronomique néo-Calédonien (IAC), Equipe Sol & Végétation (SolVeg), Nouméa, New Caledonia
- ⁸⁹Department of Forest Science, Tokyo University of Agriculture, Tokyo, Japan
- ⁹⁰Institute of Dendrology, Polish Academy of Sciences, Kórnik, Poland
- ⁹¹Department of Game Management and Forest Protection, Poznań University of Life Sciences, Poznań, Poland
- ⁹²Faculty of Biology, Białowieża Geobotanical Station, University of Warsaw, Białowieża, Poland
- ⁹³Department of Geosciences and Natural Resource Management, University of Copenhagen, Copenhagen, Denmark
- ⁹⁴Department of Plant Biology, Institute of Biology, University of Campinas, UNICAMP, Campinas, Brazil
- ⁹⁵School of Biological Sciences, University of Bristol, Bristol, UK
- ⁹⁶Forestry Faculty, Bauman Moscow State Technical University, Mytischki, Russia
- ⁹⁷CAVElab-Computational and Applied Vegetation Ecology, Department of Environment, Ghent University, Ghent, Belgium
- ⁹⁸CTFS-ForestGEO, Smithsonian Tropical Research Institute, Balboa, Panama
- ⁹⁹Department of Physical and Environmental Sciences, Colorado Mesa

- University, Grand Junction, Colorado, USA
- ¹⁰⁰UniSA STEM and Future Industries Institute, University of South Australia, Adelaide, South Australia, Australia
- ¹⁰¹Department of Botany, Dr Harisingh Gour Vishwavidyalaya (A Central University), Sagar, 470003, MP, India
- ¹⁰²Department of Agriculture, Forestry and Bioresources, Seoul National University, Seoul, South Korea
- ¹⁰³Interdisciplinary Program in Agricultural and Forest Meteorology, Seoul National University, Seoul, South Korea
- ¹⁰⁴National Center for Agro Meteorology, Seoul, South Korea
- ¹⁰⁵Research Institute for Agriculture and Life Sciences, Seoul National University, Seoul, South Korea
- ¹⁰⁶Graduate School of Agriculture, Kyoto University, Kyoto, Japan
- ¹⁰⁷Institute for World Forestry, University of Hamburg, Hamburg, Germany
- ¹⁰⁸Institute of Forestry and Rural Engineering, Estonian University of Life Sciences, Tartu, Estonia
- ¹⁰⁹Ecosystems Services and Management, International Institute for Applied Systems Analysis, Laxenburg, Austria
- ¹¹⁰Department of Geography, University College London, London, UK
- ¹¹¹Faculty of Forestry, Qingdao Agricultural University, Qingdao, China
- ¹¹²Center for Forest Ecology and Productivity, Russian Academy of Sciences, Moscow, Russia
- ¹¹³School of Geography, University of Oxford, Oxford, UK
- ¹¹⁴UMR EcoFoG, AgroParisTech, Kourou, France
- ¹¹⁵Departamento de Ciências Biológicas, Universidade do Estado de Mato Grosso, Nova Xavantina, Brazil
- ¹¹⁶Flamingo Land Ltd, Kirby Misperton, UK
- ¹¹⁷Department of Environment & Geography, University of York, York, UK
- ¹¹⁸Department of Wildlife Management, College of African Wildlife Management, Mweka, Tanzania
- ¹¹⁹Departamento de Ecología y Recursos Naturales, Facultad de Ciencias, Universidad Nacional Autónoma de México, Mexico City, Mexico
- ¹²⁰Universidad del Tolima, Ibagué, Colombia
- ¹²¹Colegio de Profesionales Forestales de Cochabamba, Cochabamba, Bolivia
- ¹²²Warsaw University of Life Sciences, Department of Forest Management, Dendrometry and Forest Economics, Warsaw, Poland
- ¹²³Jardín Botánico de Missouri, Oxapampa, Peru
- ¹²⁴Universidad Nacional de San Antonio Abad del Cusco, Cusco, Peru
- ¹²⁵Centre for Research on Land-use Sustainability, Dhaka, Bangladesh
- ¹²⁶Instituto de Silvicultura e Industria de la Madera, Universidad Juárez del Estado de Durango, Durango, Mexico
- ¹²⁷Universidad Estatal Amazónica, Puyo, Pastaza, Ecuador
- ¹²⁸The Santa Fe Institute, Santa Fe, New Mexico, USA
- ¹²⁹Department of Evolutionary Biology and Environmental Studies, University of Zürich, Zürich, Switzerland
- ¹³⁰Climate, Fire, and Carbon Cycle Sciences, USDA Forest Service, Durham, North Carolina, USA
- ¹³¹Centre for Forest Research, Université du Québec à Montréal, Montréal, Quebec, Canada
- ¹³²V. N. Sukachev Institute of Forest, FRC KSC, Siberian Branch of the Russian Academy of Sciences, Krasnoyarsk, Russia
- ¹³³Forest Ecology and Forest Management Group, Wageningen University & Research, Wageningen, The Netherlands
- ¹³⁴Department of Ecology and Environmental Sciences, Pondicherry University, Puducherry, India
- ¹³⁵Instituto Nacional de Tecnología Agropecuaria (INTA), Universidad Nacional de la Patagonia Austral (UNPA), Consejo Nacional de Investigaciones Científicas y Técnicas (CONICET), Rio Gallegos, Argentina
- ¹³⁶School of Social Sciences (Urban Studies), Western Sydney University, Penrith, New South Wales, Australia
- ¹³⁷GIP ECOFOR, Paris, France
- ¹³⁸Instituto Nacional de Pesquisas da Amazônia, Manaus, Brazil
- ¹³⁹Laboratório de Dendrologia e Silvicultura Tropical, Centro de Formação em Ciências Agroflorestais, Universidade Federal do Sul da Bahia, Itabuna, Brazil
- ¹⁴⁰Jardín Botánico de Medellín, Medellín, Colombia
- ¹⁴¹Chair for Forest Growth and Yield Science, TUM School for Life Sciences, Technical University of Munich, Munich, Germany
- ¹⁴²Universidad Nacional de la Amazonía Peruana, Iquitos, Peru
- ¹⁴³Servicios Ecosistémicos y Cambio Climático (SECC), Fundación Con Vida & Corporación COL-TREE, Medellín, Colombia
- ¹⁴⁴Centro Agricoltura, Alimenti, Ambiente, University of Trento, San Michele all'Adige, Italy
- ¹⁴⁵Department of Biological Sciences, Boise State University, Boise, Idaho, USA
- ¹⁴⁶Department of Biology, University of Florence, Florence, Italy
- ¹⁴⁷Tropical Biodiversity, MUSE—Museo delle Scienze, Trento, Italy
- ¹⁴⁸Info Flora, Geneva, Switzerland
- ¹⁴⁹Department of Environmental Sciences, Central University of Jharkhand, Ranchi, India
- ¹⁵⁰Centro de Modelación y Monitoreo de Ecosistemas, Universidad Mayor, Santiago, Chile
- ¹⁵¹Vicerrectoría de Investigación y Postgrado, Universidad de La Frontera, Temuco, Chile
- ¹⁵²Departamento de Silvicultura y Conservación de la Naturaleza, Universidad de Chile, Santiago, Chile
- ¹⁵³Faculty of Biology, Geobotany, University of Freiburg, Freiburg im Breisgau, Germany
- ¹⁵⁴Department of Geography, Remote Sensing Laboratories, University of Zurich, Zurich, Switzerland
- ¹⁵⁵National Forest Centre, Forest Research Institute Zvolen, Zvolen, Slovakia
- ¹⁵⁶Université de Lorraine, AgroParisTech, Inra, Silva, Nancy, France
- ¹⁵⁷Department of Biology, Center for Biodiversity Dynamics in a Changing World (BIOCHANGE), Aarhus University, Aarhus, Denmark
- ¹⁵⁸Faculty of Environmental Sciences and Natural Resource Management, Norwegian University of Life Sciences, Ås, Norway
- ¹⁵⁹Departamento de Biología, Universidad de la Serena, La Serena, Chile
- ¹⁶⁰Centro de Ciências Biológicas e da Natureza, Universidade Federal do Acre, Rio Branco, Acre, Brazil
- ¹⁶¹Guyana Forestry Commission, Georgetown, French Guiana
- ¹⁶²Environmental and Life Sciences, Faculty of Science, Universiti Brunei Darussalam, Bandar Seri Begawan, Brunei Darussalam
- ¹⁶³Plant Systematic and Ecology Laboratory, Department of Biology, Higher Teachers' Training College, University of Yaoundé I, Yaoundé, Cameroon
- ¹⁶⁴Departamento de Ecologia, Universidade Federal do Rio Grande do Norte, Natal, Brazil
- ¹⁶⁵Section for Ecoinformatics & Biodiversity, Department of Biology, Aarhus University, Aarhus, Denmark
- ¹⁶⁶Faculty of Forestry and Wood Sciences, Czech University of Life Sciences, Prague, Czech Republic
- ¹⁶⁷Wildlife Conservation Society, Bronx, New York, USA
- ¹⁶⁸Iwokrama International Centre for Rainforest Conservation and Development (IIC), Georgetown, French Guiana
- ¹⁶⁹Botanical Garden of Ural Branch of Russian Academy of Sciences, Ural State Forest Engineering University, Ekaterinburg, Russia
- ¹⁷⁰Pontificia Universidad Católica del Ecuador, Quito, Ecuador
- ¹⁷¹LINCGlobal, Museo Nacional de Ciencias Naturales, CSIC, Madrid, Spain
- ¹⁷²Plant Ecology and Nature Conservation Group, Wageningen University, P.O. Box 47, Wageningen, 6700 AA, The Netherlands

¹⁷³Silviculture Research Institute, Vietnamese Academy of Forest Sciences, Hanoi, Vietnam

¹⁷⁴Department of Biology, Stanford University, Stanford, California, USA

¹⁷⁵Centre for the Research and Technology of Agro-Environmental and Biological Sciences, CITAB, University of Trás-os-Montes and Alto Douro, UTAD, Vila Real, Portugal

¹⁷⁶Department of Ecology and Sustainable Agriculture, Agricultural High School of Polytechnic Institute of Viseu, Portugal and Centre for the Research and Technology of Agro-Environmental and Biological Sciences, CITAB, University of Trás-os-Montes and Alto Douro, 5001-801, Vila Real, Portugal

¹⁷⁷Department of Forest Engineering Universidade Regional de Blumenau, 89031, Blumenau, Santa Catarina, Brazil

¹⁷⁸Environmental Studies and Research Center, University of Campinas, UNICAMP, Campinas, Brazil

¹⁷⁹Department of Forest and Wood Science, University of Stellenbosch, Stellenbosch, South Africa

¹⁸⁰Key Laboratory of Tropical Biological Resources, Ministry of Education, School of Life and Pharmaceutical Sciences, Hainan University, Haikou, China

¹⁸¹Division of Forestry and Natural Resources, West Virginia University, Morgantown, West Virginia, USA

¹⁸²Department of Zoology, University of Oxford, Oxford, UK

¹⁸³Manaaki Whenua-Landcare Research, Lincoln, New Zealand

¹⁸⁴Department of Wetland Ecology, Institute for Geography and Geoecology, Karlsruhe Institute for Technology, Karlsruhe, Germany

¹⁸⁵Centre for Agricultural Research in Suriname (CELOS), Paramaribo, Suriname

¹⁸⁶Tropenbos International, Wageningen, The Netherlands

¹⁸⁷Polish State Forests, Coordination Center for Environmental Projects, Warsaw, Poland and ¹⁸⁸Research Center of Forest Management Engineering of State Forestry and Grassland Administration, Beijing Forestry University, Beijing, China

ACKNOWLEDGEMENTS

This research has been funded by a grant from DOB Ecology. Swiss National Science Foundation, Ambizione grant #PZ00P3_193612 to DSM. JCS considers this work a contribution to his VILLUM Investigator project 'Biodiversity Dynamics in a Changing World' funded by VILLUM FONDEN (grant 16549). The GFBI data from New Zealand were drawn from the Natural Forest plot data collected between January 2009 and March 2014 by the LUCAS programme for the New Zealand Ministry for the Environment and sourced from the New Zealand National Vegetation Survey Databank'. Instituto de Conservação da Natureza. FCT–UIDB/04033/2020. Open access funding provided by Eidgenössische Technische Hochschule Zurich.

CONFLICT OF INTEREST STATEMENT

The authors declare no competing interests.

PEER REVIEW

The peer review history for this article is available at <https://www.webofscience.com/api/gateway/wos/peer-review/10.1111/1365-2745.14098>.

DATA AVAILABILITY STATEMENT

The plot-level evenness data are stored in the Research Data Repository of the Research Collection of ETH Zürich <http://doi.org/10.3929/ethz-b-000597256>. The GFBI database is 3rd party data, and is publicly available upon request at <https://gfbinitiative.net/data/>.

<https://gfbinitiative.net/data/>. The GFBI database is 3rd party data, and is publicly available upon request at <https://gfbinitiative.net/data/>.

ORCID

Iris Hordijk  <https://orcid.org/0000-0002-6302-6254>

Jean-Francois Bastin  <https://orcid.org/0000-0003-2602-7247>

Helge Bruelheide  <https://orcid.org/0000-0003-3135-0356>

Roberto Cazzolla Gatti  <https://orcid.org/0000-0001-5130-8492>

Han Y. H. Chen  <https://orcid.org/0000-0001-9477-5541>

Emil Cienciala  <https://orcid.org/0000-0002-1254-4254>

David Coomes  <https://orcid.org/0000-0002-8261-2582>

Mathieu Decuyper  <https://orcid.org/0000-0002-1713-8562>

Brian Enquist  <https://orcid.org/0000-0002-6124-7096>

Tom M. Fayle  <https://orcid.org/0000-0002-1667-1189>

Henry B. Glick  <https://orcid.org/0000-0002-2956-530X>

Andrew Hector  <https://orcid.org/0000-0002-1309-7716>

Bruno Hérault  <https://orcid.org/0000-0002-6950-7286>

Andrzej M. Jagodzinski  <https://orcid.org/0000-0001-6899-0985>

Elizabeth Kearsley  <https://orcid.org/0000-0003-0046-3606>

Sebastian Kepfer-Rojas  <https://orcid.org/0000-0002-1681-2877>

Gunnar Keppel  <https://orcid.org/0000-0001-7092-6149>

Henn Korjus  <https://orcid.org/0000-0001-8522-7869>


Mait Lang  <https://orcid.org/0000-0002-0951-7933>

Huicui Lu  <https://orcid.org/0000-0001-9450-9681>

Brian Maitner  <https://orcid.org/0000-0002-2118-9880>

Jorge A. Meave  <https://orcid.org/0000-0002-6241-8803>

Sharif A. Mukul  <https://orcid.org/0000-0001-6955-2469>

Narayanaswamy Parthasarathy  <https://orcid.org/0000-0002-3445-1980>

John R. Poulsen  <https://orcid.org/0000-0002-1532-9808>

Purabi Saikia  <https://orcid.org/0000-0001-5481-282X>

Christian Salas-Eljatib  <https://orcid.org/0000-0002-8468-0829>

Peter Schall  <https://orcid.org/0000-0003-4808-818X>

Bernhard Schmid  <https://orcid.org/0000-0002-8430-3214>

Douglas Sheil  <https://orcid.org/0000-0002-1166-6591>

Krzysztof Stereńczak  <https://orcid.org/0000-0002-9556-0144>

Michael E. Van Nuland  <https://orcid.org/0000-0002-3333-0212>

Simone Vieira  <https://orcid.org/0000-0002-0129-4181>

Xiuhai Zhao  <https://orcid.org/0000-0003-0879-4063>

REFERENCES

- Aguiar, M. R., Lauenroth, W. K., & Peters, D. P. (2001). Intensity of intra- and interspecific competition in coexisting shortgrass species. *Journal of Ecology*, 89(1), 40–47.
- Ali, A., Lin, S., He, J., Kong, F., Yu, J., & Jiang, H. (2019). Climate and soils determine aboveground biomass indirectly via species diversity and stand structural complexity in tropical forests. *Forest Ecology and Management*, 432, 823–831. <https://doi.org/10.1016/j.foreco.2018.10.024>
- Balvanera, P., Pfisterer, A. B., Buchmann, N., He, J. S., Nakashizuka, T., Raffaelli, D., & Schmid, B. (2006). Quantifying the evidence for biodiversity effects on ecosystem functioning and services. *Ecology Letters*, 9(10), 1146–1156.
- Bastin, J.-F., Rutishauser, E., Kellner, J. R., Saatchi, S., Péliissier, R., Hérault, B., Slik, F., Bogaert, J., de Cannière, C., Marshall, A. R., Poulsen, J.,

- Alvarez-Loyayza, P., Andrade, A., Angbonga-Basia, A., Araujo-Murakami, A., Arroyo, L., Ayyappan, N., de Azevedo, C. P., Banki, O., ... Zebaze, D. (2018). Pan-tropical prediction of forest structure from the largest trees. *Global Ecology and Biogeography*, 27(11), 1366–1383.
- Batjes, N. H., Ribeiro, E., van Oostrum, A., Leenaars, J., Hengl, T., & Mendes de Jesus, J. (2017). WoSIS-providing standardised soil profile data for the world. *Earth System Science Data*, 9, 1–14.
- Becker, J.-M., Ringle, C. M., Sarstedt, M., & Völckner, F. (2015). How collinearity affects mixture regression results. *Marketing Letters*, 26(4), 643–659.
- Boyle, B., Hopkins, N., Lu, Z., Raygoza Garay, J. A., Mozherin, D., Rees, T., Maticsi, N., Narro, M. L., Piel, W. H., Mckay, S. J., Lowry, S., Freeland, C., Peet, R. K., & Enquist, B. J. (2013). The taxonomic name resolution service: an online tool for automated standardization of plant names. *BMC bioinformatics*, 14, 1–15.
- Brun, P., Zimmermann, N. E., Graham, C. H., Lavergne, S., Pellissier, L., Münkemüller, T., & Thuiller, W. (2019). The productivity-biodiversity relationship varies across diversity dimensions. *Nature Communications*, 10(1), 5691. <https://doi.org/10.1038/s41467-019-13678-1>
- Cardinale, B. J., Duffy, J. E., Gonzalez, A., Hooper, D. U., Perrings, C., Venail, P., Narwani, A., Mac, E. G. M., Tilman, D., Wardle, D. A., Kinzig, A. P., Daily, G. C., Loreau, M., Grace, J. B., Larigauderie, A., Srivastava, D. S., & Naeem, S. (2012). Biodiversity loss and its impact on humanity. *Nature*, 486(7401), 59–67.
- Cardinale, B. J., Matulich, K. L., Hooper, D. U., Byrnes, J. E., Duffy, E., Gamfeldt, L., Balvanera, P., O'Connor, M. I., & Gonzalez, A. (2011). The functional role of producer diversity in ecosystems. *American Journal of Botany*, 98(3), 572–592.
- Cardinale, B. J., Wright, S. J., Cadotte, M. W., Carroll, I. T., Hector, A., Srivastava, D. S., Loreau, M., & Weis, J. J. (2007). Impacts of plant diversity on biomass production increase through time because of species complementarity. *Proceedings of the National Academy of Sciences of the United States of America*, 104(46), 18123–18128.
- Caswell, H. (1976). Community structure: A neutral model analysis. *Ecological Monographs*, 46(3), 327–354.
- Ceballos, G., Ehrlich, P. R., Barnosky, A. D., García, A., Pringle, R. M., & Palmer, T. M. (2015). Accelerated modern human-induced species losses: Entering the sixth mass extinction. *Science Advances*, 1(5), e1400253. <https://doi.org/10.1126/sciadv.1400253>
- Center for International Earth Science Information Network-CIESIN-Columbia University. (2016). *Gridded Population of the World, Version 4 (GPWv4): Population Density Adjusted to Match 2015 Revision of UN WPP Country Totals*. NASA Socioeconomic Data and Applications Center (SEDAC).
- Chave, J., Condit, R., Aguilar, S., Hernandez, A., Lao, S., & Perez, R. (2004). Error propagation and scaling for tropical forest biomass estimates. *Philosophical Transactions of the Royal Society of London. Series B: Biological Sciences*, 359(1443), 409–420.
- Chave, J., Coomes, D., Jansen, S., Lewis, S. L., Swenson, N. G., & Zanne, A. E. (2009). Towards a worldwide wood economics spectrum. *Ecology Letters*, 12(4), 351–366.
- Chave, J., Réjou-Méchain, M., Búrquez, A., Chidumayo, E., Colgan, M. S., Delitti, W. B. C., Duque, A., Eid, T., Fearnside, P. M., Goodman, R. C., Henry, M., Martínez-Yrizar, A., Mugasha, W. A., Muller-Landau, H. C., Mencuccini, M., Nelson, B. W., Ngomanda, A., Nogueira, E. M., Ortiz-Malavassi, E., ... Vieilledent, G. (2014). Improved allometric models to estimate the aboveground biomass of tropical trees. *Global Change Biology*, 20(10), 3177–3190.
- Condit, R., Perez, R., Aguilar, S., Lao, S., Foster, R., & Hubbell, S. P. (2019a). *BCI 50-ha plot taxonomy, 2019 version*.
- Condit, R., Perez, R., Aguilar, S., Lao, S., Foster, R., & Hubbell, S. P. (2019b). *Complete data from the Barro Colorado 50-ha plot: 423617 trees, 35 years, 2019 version*.
- Connell, J. H. (1971). On the role of natural enemies in preventing competitive exclusion in some marine animals and in rain forest trees. In P. J. Den Boer & G. R. Gradwell (Eds.), *Dynamics of populations*. Centre for Agricultural Publishing and Documentation.
- Cook, L. M., & Graham, C. S. (1996). Evenness and species number in some moth populations. *Biological Journal of the Linnean Society*, 58, 75–84.
- Cotgreave, P., & Harvey, P. H. (1994). Evenness of abundance in bird communities. *Journal of Animal Ecology*, 63(2), 365–374.
- Dee, L. E., Cowles, J., Isbell, F., Pau, S., Gaines, S. D., & Reich, P. B. (2019). When do ecosystem services depend on rare species? *Trends in Ecology & Evolution*, 34(8), 746–758. <https://doi.org/10.1016/j.tree.2019.03.010>
- Falster, D. S., Duursma, R. A., Ishihara, M. I., Barneche, D. R., Fitz John, R. G., Vårhammar, A., Aiba, M., Ando, M., Anten, N., Aspinwall, M. J., Baltzer, J. L., Baraloto, C., Battaglia, M., Battles, J. J., Bond-Lamberty, B., van Breugel, M., Camac, J., Claveau, Y., Coll, L., ... York, R. A. (2015). BAAD: A biomass and allometry database for woody plants. *Ecology*, 96(5), 1445.
- FAO. (2000). *Global Forest Resources Assessment 2000*. FAO.
- Fei, S., Jo, I., Guo, Q., Wardle, D. A., Fang, J., Chen, A., Oswalt, C. M., & Brockerhoff, E. G. (2018). Impacts of climate on the biodiversity-productivity relationship in natural forests. *Nature Communications*, 9(1), 5436.
- Fraser, L. H., Sternberg, M., Fraser, L. H., Pither, J., Jentsch, A., Sternberg, M., Zobel, M., Askarizadeh, D., Bartha, S., Beierkuhnlein, C., & Bennett, J. A. (2015). Worldwide evidence of a unimodal relationship between productivity and plant species richness. *Science*, 349(6245), 302–306.
- Freschet, G. T., Dias, A. T. C., Ackerly, D. D., Aerts, R., van Bodegom, P. M., Cornwell, W. K., Dong, M., Kurokawa, H., Liu, G., Onipchenko, V. G., Ordoñez, J. C., Peltzer, D. A., Richardson, S. J., Shidakov, I. I., Soudzilovskaia, N. A., Tao, J., & Cornelissen, J. H. C. (2011). Global to community scale differences in the prevalence of convergent over divergent leaf trait distributions in plant assemblages. *Global Ecology and Biogeography*, 20(5), 755–765.
- GFBI Database. (2021). *Global Forest Biodiversity Initiative*. <http://www.gfbinitiative.org>
- Gitay, H., Wilson, J. B., & Lee, W. G. (1996). Species redundancy: A redundant concept? *Journal of Ecology*, 84(1), 121–124.
- Gleason, H. A. (1922). On the relation between species and area. *Ecology*, 3(2), 158–162.
- Gorelick, N., Hancher, M., Dixon, M., Ilyushchenko, S., Thau, D., & Moore, R. (2017). *Google Earth Engine: Planetary-scale geospatial analysis for everyone*. Remote Sensing of Environment.
- Grace, J. B., Anderson, T. M., Seabloom, E. W., Borer, E. T., Adler, P. B., Harpole, W. S., Hautier, Y., Hillebrand, H., Lind, E. M., Pärtel, M., Bakker, J. D., Buckley, Y. M., Crawley, M. J., Damschen, E. I., Davies, K. F., Fay, P. A., Firn, J., Gruner, D. S., Hector, A., ... Smith, M. D. (2016). Integrative modelling reveals mechanisms linking productivity and plant species richness. *Nature*, 529, 390–393.
- Grime, J. P. (1998). Benefits of plant diversity to ecosystems: Immediate, filter and founder effects. *Journal of Ecology*, 86(6), 902–910.
- Grömping, U. (2006). Relative importance for linear regression in R: The package relaimpo. *Journal of Statistical Software*, 17(1), 1–27.
- Hanlin, H., Martin, F., Wike, L., & Bennett, S. (2000). Terrestrial activity, abundance and species richness of amphibians in managed forests in South Carolina. *The American Midland Naturalist*, 143(1), 70–83.
- Hansen, M. C., Potapov, P. v., Moore, R., Hancher, M., Turubanova, S. A., Tyukavina, A., Thau, D., Stehman, S. V., Goetz, S. J., Loveland, T. R., Kommareddy, A., Egorov, A., Chini, L., Justice, C. O., & Townshend, J. R. G. (2013). High-resolution global maps of 21st-century forest cover change. *Science*, 342(6160), 850–853.
- Henry, M., Bombelli, A., Trotta, C., Alessandrini, A., Birigazzi, L., Sola, G., Vieilledent, G., Santenoise, P., Longuetaud, F., Valentini, R., Picard, N., & Saint-André, L. (2013). GlobAllomeTree: International platform for tree allometric equations to support volume, biomass and carbon assessment. *IForest-Biogeosciences and Forestry*, 6(6), 326–330.

- Hijmans, R. J., Cameron, S. E., Parra, J. L., Jones, P. G., & Jarvis, A. (2005). Very high resolution interpolated climate surfaces for global land areas. *International Journal of Climatology*, 25, 1965–1978. <https://doi.org/10.1002/joc.1276>
- Hill, M. O. (1973). Diversity and evenness: A unifying notation and its consequences. *Ecological Society of America*, 54(2), 427–432.
- Hooper, D. U., Chapin, F. S., Ewel, J. J., Hector, A., Inchausti, P., Lavorel, S., Lawton, J. H., Lodge, D. M., Loreau, M., Naeem, S., Schmid, B., Setälä, H., Symstad, A. J., Vandermeer, J., & Wardle, D. A. (2005). Effects of biodiversity on ecosystem functioning: A consensus of current knowledge. *Ecological Monographs*, 75(1), 3–35.
- Isbell, F., Gonzalez, A., Loreau, M., Cowles, J., Diaz, S., Hector, A., Mace, G. M., Wardle, D. A., O'Connor, M. I., Duffy, J. E., Turnbull, L. A., Thompson, P. L., & Larigauderie, A. (2017). Linking the influence and dependence of people on biodiversity across scales. *Nature*, 546(7656), 65–72.
- Isbell, F. I., Polley, H. W., & Wilsey, B. J. (2009). Biodiversity, productivity and the temporal stability of productivity: Patterns and processes. *Ecology Letters*, 12(5), 443–451.
- Janzen, D. H. (1970). Herbivores and the number of tree species in tropical forests. *The American Naturalist*, 104(940), 501–528. <https://doi.org/10.1086/282687>
- Jenkins, J. C., Chojnacky, D. C., Heath, L. S., & Birdsey, R. A. (2003). National-scale biomass estimators for United States tree species. *Forest Science*, 49(1), 12–35.
- Jost, L. (2010). The relation between evenness and diversity. *Diversity*, 2(2), 207–232.
- Kirwan, L., Lüscher, A., Sebastià, M. T., Finn, J. A., Collins, R. P., Porqueddu, C., Helgadottir, A., Baadshaug, O. H., Brophy, C., Coran, C., Dalmannsdóttir, S., Delgado, I., Elgersma, A., Fothergill, M., Frankow-Lindberg, B. E., Golinski, P., Grieu, P., Gustavsson, A. M., Höglind, M., ... Connolly, J. (2007). Evenness drives consistent diversity effects in intensive grassland systems across 28 European sites. *Journal of Ecology*, 95(3), 530–539.
- Lembrechts, J., de Boeck, H. J., Liao, J., Milbau, A., & Nijs, I. (2018). Effects of species evenness can be derived from species richness–Ecosystem functioning relationships. *Oikos*, 127, 337–344.
- Leys, C., Ley, C., Klein, O., Bernard, P., & Licata, L. (2013). Detecting outliers: Do not use standard deviation around the mean, use absolute deviation around the median. *Journal of Experimental Social Psychology*, 49(4), 764–766.
- Liang, J., Crowther, T. W., Picard, N., Wiser, S., Zhou, M., Alberti, G., Schulze, E. D., McGuire, A. D., Bozzato, F., Pretzsch, H., De-Miguel, S., Paquette, A., Hérault, B., Scherer-Lorenzen, M., Barrett, C. B., Glick, H. B., Hengeveld, G. M., Nabuurs, G. J., Pfautsch, S., ... Reich, P. B. (2016). Positive biodiversity–productivity relationship predominant in global forests. *Science*, 354(6309), aaf8957. <https://doi.org/10.1126/science.aaf8957>
- Lomolino, M. (2000). Ecology's most general, yet protean pattern: The species–area relationship. *Journal of Biogeography*, 27(1), 17–26.
- Loreau, M. (1998). Biodiversity and ecosystem functioning: A mechanistic model. *Proceedings of the National Academy of Sciences of the United States of America*, 95(10), 5632–5636.
- Loreau, M., & Hector, A. (2001). Partitioning selection and complementarity in biodiversity experiments. *Nature*, 412(6842), 72–76. <https://doi.org/10.1038/35083573>
- Loreau, M., Mouquet, N., & Gonzalez, A. (2003). Biodiversity as spatial insurance in heterogeneous landscapes. *Proceedings of the National Academy of Sciences of the United States of America*, 100(22), 12765–12770.
- Luo, W., Liang, J., Cazzolla Gatti, R., Zhao, X., & Zhang, C. (2019). Parameterization of biodiversity–productivity relationship and its scale dependency using georeferenced tree-level data. *Journal of Ecology*, 107(3), 1106–1119.
- Lyons, K. G., Brigham, C. A., Traut, B. H., & Schwartz, M. W. (2005). Rare species and ecosystem functioning. *Conservation Biology*, 19(4), 1019–1024.
- Ma, M. (2005). Species richness vs evenness: Independent relationship and different responses to edaphic factors. *Oikos*, 111, 192–198.
- Manier, D. J., & Hobbs, N. T. (2006). Large herbivores influence the composition and diversity of shrub-steppe communities in the Rocky Mountains, USA. *Oecologia*, 146(4), 641–651.
- McGill, B. J., Etienne, R. S., Gray, J. S., Alonso, D., Anderson, M. J., Benecha, H. K., Dornelas, M., Enquist, B. J., Green, J. L., He, F., & Hurlbert, A. H. (2007). Species abundance distributions: Moving beyond single prediction theories to integration within an ecological framework. *Ecology Letters*, 10, 995–1015.
- Nijs, I., & Roy, J. (2000). How important are species richness, species evenness and interspecific differences to productivity? A mathematical model. *Oikos*, 88(March 1999), 57–66. <https://doi.org/10.1034/j.1600-0706.2000.880107.x>
- Niklaus, P. A., Baruffol, M., He, J. S., Ma, K., & Schmid, B. (2017). Can niche plasticity promote biodiversity–productivity relationships through increased complementarity? *Ecology*, 98(4), 1104–1116. <https://doi.org/10.1002/ecy.1748>
- Olson, D. M., Dinerstein, E., Wikramanayake, E. D., Burgess, N. D., Powell, G. V. N., Underwood, E. C., D'Amico, J. A., Itoua, I., Strand, H. E., Morrison, J. C., Loucks, C. J., Allnutt, T. F., Ricketts, T. H., Kura, Y., Lamoreux, J. F., Wettengel, W. W., Hedao, P., & Kassem, K. R. (2001). Terrestrial ecoregions of the world: A new map of life on Earth: A new global map of terrestrial ecoregions provides an innovative tool for conserving biodiversity. *BioScience*, 51(11), 933–938. [https://doi.org/10.1641/0006-3568\(2001\)051\[0933:TEOTW A\]2.0.CO;2](https://doi.org/10.1641/0006-3568(2001)051[0933:TEOTW A]2.0.CO;2)
- Paquette, A., Hector, A., Castagneyrol, B., Vanhellefont, M., Koricheva, J., Scherer-Lorenzen, M., & Verheyen, K. (2018). A million and more trees for science. *Nature Ecology & Evolution*, 2(5), 763–766.
- Peet, R. K. (1974). The measurement of species diversity. *Annual Review of Ecology and Systematics*, 5(May), 285–307.
- Peichl, M., & Arain, M. A. (2006). Above- and belowground ecosystem biomass and carbon pools in an age-sequence of temperate pine plantation forests. *Agricultural and Forest Meteorology*, 140(1–4), 51–63.
- Poorter, L., Bongers, F., Aide, T. M., Almeida Zambrano, A. M., Balvanera, P., Becknell, J. M., Boukili, V., Brancalion, P. H. S., Broadbent, E. N., Chazdon, R. L., Craven, D., de Almeida-Cortez, J. S., Cabral, G. A. L., de Jong, B. H. J., Denslow, J. S., Dent, D. H., DeWalt, S. J., Dupuy, J. M., Durán, S. M., ... Rozendaal, D. M. A. (2016). Biomass resilience of neotropical secondary forests. *Nature*, 530(7589), 211–214.
- Poulter, B., Aragão, L., Andela, N., Bellassen, V., Ciais, P., Kato, T., Lin, X., Nachin, B., Luyssaert, S., Pederson, N., Peylin, P., Piao, S., Pugh, T., Saatchi, S., Schepaschenko, D., Schelhaas, M., & Shvidenko, A. (2019). *The global forest age dataset and its uncertainties (GFADv1.1)*. NASA National Aeronautics and Space Administration.
- R Core Team. (2018). *R: A language and environment for statistical computing*. R Foundation for Statistical Computing. <https://cran.r-project.org/package=vegan>
- Réjou-Méchain, M., Tanguy, A., Pioniot, C., Chave, J., & Hérault, B. (2017). Biomass: An r package for estimating above-ground biomass and its uncertainty in tropical forests. *Methods in Ecology and Evolution*, 8(9), 1163–1167.
- Ribeiro, E., Batjes, N., & van Oostrum, A. (2018). *World Soil Information Service (WoSIS)—Towards the standardization and harmonization of world soil data. Procedures Manual 2018*. ISRIC report 2018/01, ISRIC-World Soil Information.
- Running, S., Mu, Q., & Zhao, M. (2011). *MOD17A3 MODIS/Terra Net Primary Production Yearly L4 Global 1km SIN Grid V055 [Data set]*. NASA EOSDIS Land Processes DAAC.

- Schmid, B. (2002). The species richness-productivity controversy. *Trends in Ecology & Evolution*, 17(3), 113–114.
- Sheil, D., & Bongers, F. (2020). Interpreting forest diversity-productivity relationships: Volume values, disturbance histories and alternative inferences. *Forest Ecosystems*, 7(1), 6. <https://doi.org/10.1186/s40663-020-0215-x>
- Slik, J. W. F., Paoli, G., McGuire, K., Amaral, I., Barroso, J., Bastian, M., Blanc, L., Bongers, F., Boundja, P., Clark, C., Collins, M., Dauby, G., Ding, Y., Doucet, J.-L., Eler, E., Ferreira, L., Forshed, O., Fredriksson, G., Gillet, J.-F., ... Zweifel, N. (2013). Large trees drive forest aboveground biomass variation in moist lowland forests across the tropics. *Global Ecology and Biogeography*, 22(12), 1261–1271.
- Soininen, J., Passy, S., & Hillebrand, H. (2012). The relationship between species richness and evenness: A meta-analysis of studies across aquatic ecosystems. *Oecologia*, 169, 803–809.
- Sonkoly, J., Kelemen, A., Valkó, O., Deák, B., Kiss, R., Tóth, K., Migléc, T., Tóthmérész, B., & Török, P. (2019). Both mass ratio effects and community diversity drive biomass production in a grassland experiment. *Scientific Reports*, 9(1), 1848.
- Spake, R., Mori, A. S., Beckmann, M., Martin, P. A., Christie, A. P., Duguid, M. C., & Doncaster, C. P. (2021). Implications of scale dependence for cross-study syntheses of biodiversity differences. *Ecology Letters*, 24(2), 374–390.
- Stirling, G., & Wilsey, B. (2001). Empirical relationships between species richness, evenness, and proportional diversity. *The American Naturalist*, 158(3), 286–299.
- Symonds, M. R. E., & Johnson, C. N. (2008). Species richness and evenness in Australian birds. *The American Naturalist*, 171(4), 480–490.
- ter Steege, H., Pitman, N. C. A., Sabatier, D., Baraloto, C., Salomão, R. P., Guevara, J. E., Phillips, O. L., Castilho, C. v., Magnusson, W. E., Molino, J. F., Monteagudo, A., Vargas, P. N., Montero, J. C., Feldpausch, T. R., Coronado, E. N. H., Killeen, T. J., Mostacedo, B., Vasquez, R., Assis, R. L., ... Silman, M. R. (2013). Hyperdominance in the Amazonian tree flora. *Science*, 342(6156), 1243092. <https://doi.org/10.1126/science.1243092>
- Tramer, E. J. (1969). Bird species diversity: Components of Shannon's formula. *Ecology*, 50(5), 927–929.
- Tuanmu, M.-N., & Jetz, W. (2014). A global 1-km consensus land-cover product for biodiversity and ecosystem modelling. *Global Ecology and Biogeography*, 23(9), 1031–1045.
- Tuomisto, H. (2012). An updated consumer's guide to evenness and related indices. *Oikos*, 121, 1203–1218.
- van Ruijven, J., & Berendse, F. (2005). Diversity-productivity relationships: Initial effects, long-term patterns, and underlying mechanisms. *Proceedings of the National Academy of Sciences of the United States of America*, 102(3), 695–700.
- Viana, D. S., & Chase, J. M. (2019). Spatial scale modulates the inference of metacommunity assembly processes. *Ecology*, 100(2), e02576.
- Wilson, J. B., Steel, J. B., McG. King, W., & Gitay, H. (1999). The effect of spatial scale on evenness. *Journal of Vegetation Science*, 10(4), 463–468.
- Yachi, S., & Loreau, M. (1999). Biodiversity and ecosystem productivity in a fluctuating environment: The insurance hypothesis. *Proceedings of the National Academy of Sciences of the United States of America*, 96(4), 1463–1468.
- Yan, Y., Connolly, J., Liang, M., Jiang, L., & Wang, S. (2021). Mechanistic links between biodiversity effects on ecosystem functioning and stability in a multi-site grassland experiment. *Journal of Ecology*, 109(9), 3370–3378.
- Zhang, H., John, R., Peng, Z., Yuan, J., Chu, C., Du, G., & Zhou, S. (2012). The relationship between species richness and evenness in plant communities along a successional gradient: A study from sub-alpine meadows of the eastern Qinghai-Tibetan plateau, China. *PLoS ONE*, 7(11), 1–9. <https://doi.org/10.1371/journal.pone.0049024>
- Zhang, Y., Chen, H. Y. H., & Reich, P. B. (2012). Forest productivity increases with evenness, species richness and trait variation: A global meta-analysis. *Journal of Ecology*, 100(3), 742–749.

SUPPORTING INFORMATION

Additional supporting information can be found online in the Supporting Information section at the end of this article.

Figure S1: The relationship between evenness and richness is more negative, than a null model comprised of 10,000 random subsets of the global GFBI data.

Figure S2: The global relationship between evenness and logged species richness for several commonly used evenness indices ($N=20,272$) (Tuomisto, 2012).

Figure S3: DBH-based allometric models of tree biomass for non-tropical biomes based on the approach presented in Jenkins et al. (2003).

Figure S4: The Pearson correlation coefficient of the relationship between Hill's evenness and the log of species richness for the boreal, temperate and tropical biomes.

Figure S5: Visualising the differences between small, medium and large plot sizes on (A) the relationship between Hill's evenness and richness, (B) the interaction between evenness, richness and Modis NPP at four different levels of evenness, and (C) the interaction between evenness, richness and biomass accumulation at four different levels of evenness.

Figure S6: Analysis of the sensitivity to plot size of a hypothetical relationship among species richness, evenness and their interaction when predicting NPP or biomass accumulation rates (A) in the global dataset and (B–D) on biome level.

Figure S7: The hypothesized effect of different levels of evenness for the global relationship between species richness and Modis NPP.

Figure S8: The global relationship between evenness, richness and (A) biomass accumulation, or (C) Modis derived NPP, with uncertainty incorporated.

Table S1: Allometric equations used for each of the non-tropical biomes, following Jenkins et al. (2003).

Table S2: Overview of the WWF biomes.

How to cite this article: Hordijk, I., Maynard, D. S., Hart, S. P., Lidong, M., ter Steege, H., Liang, J., de-Miguel, S., Nabuurs, G.-J., Reich, P. B., Abegg, M., Adou Yao, C. Y., Alberti, G., Almeyda Zambrano, A. M., Alvarado, B. V., Esteban, A.-D., Alvarez-Loayza, P., Alves, L. F., Ammer, C., Antón-Fernández, C. ... Crowther, T. W. (2023). Evenness mediates the global relationship between forest productivity and richness. *Journal of Ecology*, 111, 1308–1326. <https://doi.org/10.1111/1365-2745.14098>